

1 **Glyphosate-based herbicide has soil-mediated effects on potato**
2 **glycoalkaloids and oxidative status of a potato pest**

3
4 Miia J. Rainio^{a*}, Aigi Margus^b, Valtteri Virtanen^c, Leena Lindström^b, Juha-Pekka Salminen^c, Kari
5 Saikkonen^d, Marjo Helander^a

6
7 ^a*Department of Biology, University of Turku, FI-20014 TURKU, FINLAND (email: Miia Rainio:*
8 *miikoi@utu.fi, Marjo Helander: helander@utu.fi)*

9 ^b*Department of Biological and Environmental Science, University of Jyväskylä, FI-40014*
10 *JYVÄSKYLÄ, FINLAND (email: Aigi Margus: aigi.margus@jyu.fi, Leena Lindström:*
11 *leena.m.lindstrom@jyu.fi)*

12 ^c*Department of Chemistry, University of Turku, FI-20014 TURKU, FINLAND (email: Juha-Pekka*
13 *Salminen: j-p.salminen@utu.fi, Valtteri Virtanen: valtteri.virtanen@utu.fi)*

14 ^d*Biodiversity Unit, University of Turku, FI-20014 TURKU, FINLAND (email: Kari Saikkonen:*
15 *karisaik@utu.fi)*

16
17 *Corresponding author: Miia Johanna Rainio, Department of Biology, University of Turku, FI-
18 20014 Turku, Finland. Tel.: +358 2 333 6050; Fax: + 358 2 333 6550; Email: miikoi@utu.fi

19
20
21 **Abbreviations**

22 CAT = catalase, GBH = glyphosate-based herbicide, GP = glutathione peroxidase, GR = glutathione reductase, GSH =
23 glutathione, GSH:GSSG = reduced vs. oxidized form of glutathione, GST = glutathione-S-transferase, LHP = lipid
24 hydroperoxides, ROS = reactive oxygen species, SOD = superoxide dismutase, tGSH = total glutathione

25

26 **Highlights**

27

28 The α -solanine levels were reduced in potato plants grown in GBH-treated soil.

29

30 The survival of the beetles was not affected by the soil-mediated GBH treatment.

31

32 Indirect GBH treatment modify the antioxidant defense of the Colorado potato beetle larvae.

33

34 Soil-mediated GBH treatment at larval stage may have long-term effects on the adult beetles.

35

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

51 **Abstract**

52

53 Glyphosate is the most used herbicide worldwide, targeting physiological pathways in plants. Recent
54 studies have shown that glyphosate can also cause toxic effects in animals. We investigated the
55 glyphosate-based herbicide (GBH)-induced changes in potato (*Solanum tuberosum*) plant chemistry
56 and the effects of a GBH on the survival rate and oxidative status of the Colorado potato beetle
57 (*Leptinotarsa decemlineata*). The beetles were reared on potato plants grown in pots containing soil
58 treated with a GBH (Roundup Gold, 450 g/l) or untreated soil (water control). The 2nd instar larvae
59 were introduced to the potato plants and then collected in 2 phases: as 4th instar larvae and as adults.
60 The main glycoalkaloids of the potato plants, α -solanine and α -chaconine, were measured twice
61 during the experiment. The α -solanine was reduced in potato plants grown in GBH-treated soil, which
62 can be detrimental to plant defenses against herbivores. GBH treatment had no effect on the survival
63 rate or body mass of the larvae or the adult beetles. In the larvae, total glutathione (tGSH)
64 concentration and the enzyme activity of catalase (CAT), superoxide dismutase, and glutathione-S-
65 transferase were increased in the GBH treatment group. In the adult beetles, CAT activity and tGSH
66 levels were affected by the interactive effect of GBH treatment and the body mass. To conclude,
67 environmentally relevant concentrations of a GBH can affect the potato plant's glycoalkaloid
68 concentrations, but are not likely to directly affect the survival rate of the Colorado potato beetle, but
69 instead, modify the antioxidant defense of the beetles via diet.

70

71

72

73 **Keywords:** Antioxidant defense, Herbivores, Insects, Potato defense chemicals, Roundup, α -solanine

74

75

76 **1. Introduction**

77

78 Glyphosate (N-(phosphonomethyl)glycine) is the most commonly used herbicide worldwide, given
79 its effectiveness and broad spectrum ability to kill weeds (Myers et al., 2016; Woodburn, 2000). It
80 has been proclaimed to be safe for the environment due to its low accumulation rate and rapid
81 inactivation in soils (Giesy et al., 2000, Vereecken, 2005). However, accumulating evidence has
82 demonstrated that glyphosate and its degradation metabolites (e.g., aminomethylphosphonic acid,
83 AMPA) can remain in the soil for years and affect non-target organisms (Helander et al., 2018; Larsen
84 et al., 2012). Furthermore, non-target organisms may be directly exposed to glyphosate products by
85 the unwanted loss of substance during transportation, handling, and storage, and by wind action
86 during field application (Torretta et al., 2018). Glyphosate exposure may also occur when it is used
87 to synchronize and accelerate the ripening of forage cereals (Helander et al., 2012). Glyphosate use
88 is intended to tackle weeds, but recent toxicological studies have shown harmful effects of glyphosate
89 products in animals, such as changes in cell function, tissues, physiology, and survival rate of the
90 animals (Claus et al., 2016; Margus et al., 2019; Mesnage et al., 2015).

91

92 Glyphosate is also the most important herbicide directly affecting the synthesis of secondary
93 compounds in plants (Duke and Powles, 2008). The glyphosate-based reduction of secondary
94 compounds in plants (i.e., defense chemicals) may expose plants to herbivore attacks; influence the
95 flavor-producing chemicals important in herbivore behavior or food quality (El-keltawi and Croteau,
96 1987); and reduce plant resistance to pathogens and fungal infections (Lydon and Duke, 1989). On
97 the other hand, glyphosate may also increase the production of plant secondary metabolites (Ossipov
98 et al. 2003). Overall, the sub-lethal effects of herbicides on non-target plants may affect agricultural
99 ecosystems by altering the synthesis of compounds that are important in inter- and intraspecific
100 interactions (Lydon and Duke, 1989). Plant-herbivore interactions are central to both food production

101 and biological diversity, affecting the dynamics of various ecosystems (Blumenthal and Augustine,
102 2009).

103

104 Glyphosate is the only herbicide affecting the inactivation of the 5-enolpyruvylshikimate-3-phosphate
105 synthase (EPSPS) enzyme (Duke and Powles, 2008; Steinrücken and Amrhein, 1980). This enzyme
106 belongs to the shikimate metabolic pathway, which appears in plants and in some bacteria and fungi
107 (Bentley, 1990; Haslam, 1993; Helander et al., 2018). Glyphosate blocks phosphoenolpyruvate (PEP)
108 binding sites, thus inhibiting the reaction between shikimate 3-phosphate (S3P) and PEP (Funke et
109 al., 2006). An inactivation of EPSPS leads to the accumulation of high levels of shikimate in plant
110 tissues (Amrhein et al., 1980; Lydon and Duke, 1989), preventing the biosynthesis of essential
111 aromatic amino acids (e.g., phenylalanine, tyrosine, and tryptophan) necessary in protein synthesis
112 (Duke and Powles, 2008) and as precursors for several secondary metabolites important in plant
113 growth (Tzin and Galili, 2010). This can result in shortages of carbon for other essential pathways
114 (Siehl, 1997) and reduce (Kishore and Shah, 1988; Martinez et al., 2018; Shilo et al., 2016; Sihtmäe
115 et al., 2013) or increase (Ossipov et al., 2003) secondary metabolites in some species of plants and
116 microbes. For example, while blocking the production of arogenic acid, glyphosate may direct the
117 conversion of secondary metabolites into hydrolysable tannins via 3-dehydroshikimic acid, which
118 have been shown to accumulate under glyphosate treatment (Ossipov et al., 2003). Glyphosate is also
119 a strong chelating agent that creates the complexes that immobilize the mineral micronutrients of soil,
120 making them unavailable to plants (Glass, 1984).

121

122 Both glyphosate and plant defense chemicals are known to impair the antioxidant defense system and
123 increase the production of reactive oxygen species (ROS) in plants (Adamski et al., 2014; Chowański
124 et al., 2016; Gomes et al., 2016; Liu et al., 2010, Radman and Fayeze, 2016) and animals (Annett et
125 al., 2014; Hultberg, 2007; Modesto and Martinez, 2010; Uren Webster and Santos, 2015), which can,

126 in turn, cause cellular biochemical stress, called oxidative stress, and consequent oxidative damage
127 to biomolecules (George and Gatehouse, 2013; Halliwell and Gutteridge, 2007). Previous studies in
128 animals have shown increased oxidative stress or alteration in antioxidant defense systems in relation
129 to various glyphosate-based herbicides (thereafter GBHs; Contardo-Jara et al., 2009; El-Shenawy,
130 2009; Gluszcak et al., 2007; Modesto and Martinez, 2010; Rainio et al., 2019; Uren Webster and
131 Santos, 2015). Also, the breakdown products of glutathione (e.g. γ -glutamylglutamine and
132 cysteinylglycine), involved in the regulation of redox balance, have been shown to increase in rats
133 exposed to GBH (Mesnage et al. 2019). Moreover, GBHs have been shown to affect the survival rate,
134 development, and reproduction of invertebrates found in agroecosystems (Benamú et al., 2010;
135 Castilla et al., 2010; Evans et al., 2010; Saska et al., 2016; Schneider et al., 2009), though there are
136 also studies reporting little or no effects (Margus et al., 2019; Salvio et al., 2016; Thompson et al.,
137 2014). The impacts of GBHs on plants and non-target organisms may differ substantially depending
138 on the use of commercial formulations that differ in their surfactant and salts, which are added to
139 enhance the effectiveness of glyphosate. Some adjuvants used in GBHs may be even more toxic than
140 the glyphosate itself (Mesnage et al., 2014). Previous studies have shown that the consequences of
141 GBH use in target ecosystems and their surrounding areas are relatively poorly known and require
142 further studies from a multidisciplinary approach.

143

144 The increasing evidence of glyphosate toxicity on non-target organisms has caused growing concern
145 about the use of glyphosate as the primary weed management strategy (Helander et al., 2012; Torretta
146 et al., 2018; Van Bruggen et al., 2018). The environmental risks of glyphosate are likely to be
147 pronounced in northern ecosystems, which are characterized by long biologically inactive winters
148 and short growing seasons, limiting the time period of peak glyphosate degradation activity to the
149 summer months (Laitinen, 2009; Helander et al., 2012; Helander et al., 2018; Silva et al., 2018). On

150 the other hand, plant-protective agents are required for effective crop production, thus it is important
151 to find safe and sustainable ways to protect plants in the future.

152

153 In this study, we investigated the soil-mediated effects of a GBH on the glyphosate-induced changes
154 in plant chemistry, and the survival rate and oxidative status of a non-target herbivore, by using potato
155 plant (*Solanum tuberosum*) and the Colorado potato beetle (*Leptinotarsa decemlineata*, Coleoptera,
156 Chrysomelidae) as a model system. The Colorado potato beetle is an economically important potato
157 pest worldwide (Casagrande, 1987; Grapputo et al., 2005; Walsh, 1865;), including in Finland, where
158 it is classified as a quarantine pest species (Vänninen et al., 2011). Potato plants and the Colorado
159 potato beetle form an excellent study system, since glyphosate is known to affect herbivores not only
160 directly, but also via potato plant defense chemicals. At the larval stage, the beetles can be exposed
161 to glyphosate residues or glyphosate metabolites via diet or due to possible changes in potato plant
162 quality; whereas, at the pupal stage, the beetles may be exposed to GBH residues also via the soil.

163

164 Potato plants are characterized by the presence of steroidal glycoalkaloids, such as α -solanine and α -
165 chaconine (Lachman et al., 2001; Matthews et al., 2005), which are biosynthetically derived from
166 cholesterol (Chowański et al., 2016). These glycoalkaloids are produced in all parts of the plant,
167 having the highest concentrations in the leaves, flowers, and unripe fruits (Adamski et al., 2014;
168 Friedman, 2006). Glycoalkaloids have insecticidal and fungicidal properties, and are often
169 synthesized when plants are under stress, such as when they have been injured by herbivores
170 (Chowański et al., 2016). They disrupt the cellular functions of herbivores, increase the generation of
171 ROS (Chowański et al., 2016), act as acetylcholinesterase inhibitors (Friedman et al., 1997), and also
172 elicit behavioral responses by insects (Lyytinen et al., 2007; Nylin and Janz, 1993). Potato plant
173 glycoalkaloids have been previously shown to reduce the growth rate and food consumption rate in
174 the khapra beetle (*Trogoderma granarium*; Nenaah, 2011), decrease reproduction rates in the potato

175 aphid (*Macrosiphum euphorbiae*; Güntner et al., 1997); decrease fertility, survival rate, and
176 hatchability in the greater wax moth (*Galleria mellonella*; Adamski et al., 2014); and increase
177 mortality in peach potato aphids (*Myzus persicae*; Fragoyiannis et al., 1998). On the other hand, it is
178 possible that under a certain threshold level of foliage glycoalkaloids, the herbivores may still feed
179 and reproduce (Khan et al., 2013). Colorado potato beetle larvae have shown either negative (Hare,
180 1987) or no response (Kowalski et al., 1999) in relation to glycoalkaloids, suggesting that the effects
181 of glycoalkaloids may vary with the life-stage of the beetle or the length of exposure (Lyytinen et al.,
182 2007).

183

184 To examine the soil-mediated effects of the GBH on the oxidative status of the beetles, we measured
185 antioxidant glutathione (total glutathione, tGSH) and the ratio of its reduced and oxidized form
186 (GSH:GSSG). Glutathione (GSH) is one of the most important small antioxidant molecules in almost
187 all organisms (Andrews, 2000) and the GSH:GSSG ratio, which indicates the overall redox status of
188 cells, is commonly used as an indicator of oxidative stress (Halliwell and Gutteridge, 2007; Isaksson
189 et al., 2005; Rainio et al., 2013). In addition, we measured the activity of insect homologs' antioxidant
190 enzymes glutathione peroxidase (GPx) and glutathione reductase (GR), as well as glutathione-S-
191 transferases (GSTs) related to GSH metabolism (Halliwell and Gutteridge, 2007). GSTs are a
192 ubiquitous and important family of enzymes (isozymes) participating in detoxification processes by
193 catalyzing the conjugation of GSHs on xenobiotics (Alghamdi and Frey, 2017; Halliwell and
194 Gutteridge, 2007) and showing the peroxidative activity function in insects (Corona and Robinson,
195 2006; Farjan et al., 2012). ROS regulation enzymes, superoxide dismutase (SOD) and catalase (CAT),
196 were measured to study first-line antioxidant defense (Fridovich, 1974), where superoxides are
197 transformed to hydrogen peroxide (H₂O₂) by SOD and further catalyzed to water (H₂O) and molecular
198 oxygen by CAT (Finkel and Holbrook, 2000; Pinto et al., 2003). To determine oxidative damage, we
199 measured lipid hydroperoxides (LHP), which have been suggested to increase with ROS production.

200 Lipid peroxidation can be harmful in insects, because, in addition to being essential components in
201 cell membranes, they also have unique physiological functions (e.g., in developmental and
202 reproductive physiology; Downer, 1985).

203

204 We hypothesize the following: **1)** Environmentally relevant levels of a GBH in the soil may cause
205 quantitative effects in the production of glycoalkaloids, since GBHs affect the aromatic amino acid
206 L-tryptophan (Santos-Sánchez et al., 2019), which is a precursor of alkaloids in secondary
207 metabolism (Dewick, 2009). If the GBH affects plant defense chemicals, it may change the plant
208 quality and resource allocation for growth and defense and change plant-herbivore interactions by
209 making the potato plants more (or less) sensitive to herbivore attacks. **2)** The GBH may reduce the
210 survival rate and body mass of the beetle larvae and adult beetles, and increase the developmental
211 time of the adult beetles in cases where the GBH is absorbed into the potato plant via the soil. **3)** The
212 GBH may further show negative soil-mediated effects during the pupal stage of the beetles, which
213 may reflect the adult's survival rate as well. **4)** The GBH may affect the antioxidant defense system
214 of the beetles by changing the antioxidant enzyme activity or GSH concentrations, either via diet or
215 soil-mediated effects during the pupal stage of the beetles.

216

217 **2. Materials and methods**

218

219 *2.1. Study design*

220

221 The GBH (Roundup Gold, Monsanto, USA) treatment was conducted in summer 2016 in a licensed
222 quarantine greenhouse in the Botanical Garden of the University of Turku (60° 26' N, 22°10' E). We
223 preferred to use the commercial formulation of glyphosate rather than pure glyphosate, since those
224 are more relevant in the agricultural context. To study the soil-mediated effects of the GBH on the

225 Colorado potato beetles in the greenhouse experiment, we used soil that had been pre-treated with the
226 GBH. The soil was collected from a long-term field experiment established in 2013 at the Botanical
227 Garden (see more details in Hagner et al., 2019). The experimental soil was treated with a permitted
228 dose of Roundup Gold (450 g/l isopropylamine glyphosate salt, CAS: 38641-94-0, application rate:
229 6.4 l/ha) that was applied twice per year (specifically, May 2014, 2015, and 2016; and October 2014
230 and 2015). The control soil received the same amount of tap water as the treated soil. The soil type
231 in the field was medium clay with a high organic matter content ($>120 \text{ g kg}^{-1}$) and pH 5.9. In June
232 2016, the soil for the greenhouse experiment was collected from the field experiment 2 weeks after a
233 GBH treatment and divided into 100 pots (\varnothing 19 cm; 50 controls, 50 GBH-treated). The organic variety
234 ‘Ditta’ potatoes were planted in the pots with the GBH-treated and control soils, and the pots were
235 then fully randomized in the greenhouse. The position of the pots was further changed during the
236 growing period to prevent the uneven growth of the potato plants. The plants were grown in ambient
237 June-July day-lengths in southwest Finland (about 17-19 h day length) under a 20°C/15°C day/night
238 temperature.

239

240 We used the United States (Vermont) Colorado potato beetle population collected from the field
241 (44°43'N, 73°20'W) in 2010, which had been since grown in laboratory conditions (see Lehmann et
242 al., 2015). Altogether, 500 Colorado potato beetle larvae (250 larvae/treatment group, 30
243 larvae/family) from 16 families (full-sib design) were used in this experiment. After 3.5 weeks of the
244 potato planting, small 2-day-old larvae (2st instar) were randomly introduced to the potato plants (5
245 larvae to each plant), which were covered by light-permeable fabric bags. After 9 days, when the
246 larvae were at their 4th instar, 184 larvae (94 controls, 90 GBH-treated) were collected, weighed, and
247 stored in a freezer at -80°C for oxidative status analyses. The remaining larvae were grown until they
248 dropped from the plant and burrowed into the soil to pupate. Once all larvae had burrowed into the
249 soil, the potato plant shoots were cut and removed. Emerged adult beetles (133 controls, 134 GBH-

250 treated) were collected every day, weighed, sexed, and used for oxidative status analyses to study the
251 possible soil-mediated or carry-over effects of the GBH. To analyze potato plant glycoalkaloids,
252 α -solanine and α -chaconine, we took ca 5 leaves per potato plant a) before placing the larvae on the
253 plants (1st measurement) and b) when the larvae had pupated and the shoots had been cut down (2nd
254 measurement). Leaves were freeze-dried, ground (TissueLyser, Qiagen, Austin, TX, USA), and
255 stored in a freezer at -20°C until the chemical analyses. The licenses for rearing quarantine pest
256 species in laboratory conditions were given by the Finnish Food Authority, Finland (Ruokavirasto,
257 permission 4057/0614/2016). Licenses for conducting experiments with insects are not necessary in
258 Finland.

259

260 *2.2. Determination of potato plant defense chemicals*

261

262 For the quantitation of potato plant glycoalkaloids, α -chaconine and α -solanine, 5 mg of ground
263 potato plant leaf material was weighed in a 2 ml Eppendorf tube. Samples were extracted with 2 ml
264 of 5% aqueous acetic acid (5:95, v/v) utilizing overnight maceration in a cold room (4°C) and were
265 shaken with a planar shaker (280 min⁻¹) for 3 hours at room temperature. Extracts were centrifuged
266 (14,000 min⁻¹) for 10 min and decanted into new 2 ml Eppendorf tubes. 100 × dilutions were made
267 with the extraction solvent and samples were filtered via polytetrafluoroethylene filters (13 mm i.d.;
268 0.2 μ m) and analyzed with a UHPLC-DAD-ESI-Orbitrap-MS instrument. One of the potato plant leaf
269 extracts was chosen as the quality control sample. It was analyzed before and after every 10 samples
270 to monitor the changes in the performance of the mass spectrometer. The ultrahigh performance liquid
271 chromatograph was coupled to a photodiode array detector (UHPLC-DAD, Waters Corporation,
272 Milford, MA, USA) and a hybrid quadrupole-Orbitrap mass spectrometer (Q Exactive, Thermo Fisher
273 Scientific, Bremen, Germany). ACQUITY UPLC BEH Phenyl (100*2.1 mm i.d., 1.7 μ m, Waters
274 Technologies Ireland, Wexford, Ireland) columns were utilized. The mobile phase consisted of

275 acetonitrile (A) and 0.1% aqueous formic acid (99.9:0.1, v/v) (B): 0-0.5 min, 0.1% A in B; 0.5-6 min,
276 0.1-30% A in B; and 6-10.5 min, column wash and stabilization. The heated electrospray ionization
277 (ESI) source (H-ESI, Thermo Fisher Scientific, Bremen, Germany) was operated in the positive ion
278 mode. Source parameters were as follows: spray voltage, +3.8 kV; sheath gas (N₂) flow rate, 60
279 (arbitrary units); auxiliary gas (N₂) flow rate, 20 (arbitrary units); sweep gas flow rate, 0 (arbitrary
280 units); capillary temperature, 380°C. The Orbitrap spectrometer was operated with a resolution of
281 35,000 and a mass range of m/z 150-2250. Data processing was done using Thermo Xcalibur Quan
282 Browser software (Version 4.1.31.9, Thermo Fisher Scientific, Waltham, MA, USA). Concentrations
283 of α -chaconine and α -solanine in samples were quantified using external calibration curves made
284 from the commercial standards of both α -chaconine and α -solanine (Carbosynth, Compton, UK).

285

286 *2.3. Oxidative status analyses*

287

288 Beetle homogenates (larvae and adults) were used to measure oxidative status biomarkers (GST, GPx,
289 GR, CAT, SOD, tGSH, and GSH:GSSG) and oxidative damage (LHP) of the beetles. All antioxidant
290 and enzyme activities was measured in triplicate (intra-assay coefficient of variability [CV] < 15% in
291 all cases) using 96- (CAT and LHP) or 384-well (GPx, GR, GST, SOD, tGSH, and GSH:GSSG)
292 microplates, which in most cases required reducing the reagent volumes as per the kit instructions.
293 All analyses were measured with an EnVision[®] microplate reader (PerkinElmer Finland, Turku,
294 Finland). There were 3 control samples used with each plate to be able to correct inter-assay precision
295 with the ratio specific to the particular plate (range 0.8-1.2).

296

297 Samples were homogenized individually (TissueLyser, Qiagen, Austin, TX, USA) with 180 μ l
298 (larvae) or 150 μ l (adults) KF buffer (0.1 M K₂HPO₄ + 0.15 M KCl, pH 7.4). The protein
299 concentration (mg/ml) was measured with bicinchoninic acid (BCA) protein assay (Smith et al., 1985)

300 using bovine serum albumin (BSA) as a standard (Sigma-Aldrich Finland, Espoo, Finland) with an
301 EnVision® microplate reader at an absorbance of 570 nm.

302

303 GST assay (Sigma-Aldrich CS0410) was adjusted from a 96- to 384-well plate. We used 2 µl of each
304 sample in triplicate and our own reagents: Dulbecco's phosphate-buffered saline (DPBS), 200 mM
305 GSH (Sigma G4251), and 100 mM 1-Chloro-2,4-dinitrobenzene (CDNB; Sigma-Aldrich C6396) in
306 ethanol. The change in absorbance was measured at 340 nm. GPx assay (Sigma-Aldrich CGP1) was
307 adjusted from a cuvette to a 384-well plate and the activity was measured according to kit instructions,
308 using 2 mM H₂O₂ instead of t-Bu-OOH as a substrate (see details in Rainio et al., 2019). The change
309 in absorbance was measured at 340 nm. GR-assay (Sigma-Aldrich GR-SA) was adjusted from a
310 cuvette to a 384-well plate and modified from the kit instructions by using our own reagents: assay
311 buffer (100 mM potassiumphosphate buffer + 1 mM EDTA, pH 7.5), 2 mM GSSG (Sigma-Aldrich
312 GG4626), 3 mM DTNB (Sigma-Aldrich D8130), and 2 mM NADPH (Sigma-Aldrich N1630). The
313 change in absorbance was measured at 412 nm. SOD assay (Sigma-Aldrich 19160) was adjusted from
314 96- to 384-well plate and measured according to kit instructions. We used 0.3 mg/ml sample dilution
315 and the activity was expressed as inhibition % at an absorbance of 450 nm. CAT-assay (Sigma-
316 Aldrich CAT100) was adjusted from a cuvette to a 96-well plate. We used 0.6 mg/ml sample dilution
317 and tested each sample in triplicate. We made our own reagents: 10 × CAT assay buffer (500 mM
318 KF, pH 7.0), CAT dilution buffer (50 mM KF + 0.1% TritonX, pH 7.0), chromogen reagent (0.25
319 mM 4-aminoantipyrene + 2 mM 3,5-dicloro-2-hydroxybenzenesulfonic acid in 150 mM potassium
320 phosphate buffer, pH 7.0), peroxidase solutions (from horseradish), stop solution (15 mM NaN₃,
321 Sigma-Aldrich), and 200 mM and 10 mM H₂O₂ according to information provided in the technical
322 bulletin (see also Deisseroth and Dounce, 1970; Fossati et al., 1980). The change in absorbance was
323 measured at 520 nm. Total GSH and the ratio of GSH:GSSG were measured with a ThioStar®
324 Glutathione Fluorescent Detection Kit (K005-FI, Arbor Assays, Ann Arbor, MI, USA) according to

325 kit instructions, and the fluorescence was measured at an excitation/emission wavelength of
326 405/510 nm. Prior to analyses, the sample homogenate was deproteinized with 5% sulfosalicylic acid
327 (SSA), incubated on ice for 10 min, and centrifuged for 10 min at 10,000 g in 4°C.

328

329 For the LHP measurement, the larvae were first weighed and then homogenized with 125 μ l methanol.
330 LHP were measured using the FOX-II method, modified from Nourooz-Zadeh et al. (1995) and Bou
331 et al. (2008). We used 45 μ l of the sample, 5 μ l 10 mM thiamine pyrophosphate (TPP) or methanol,
332 and 950 μ l of FOX reagent (see also Vuori et al., 2015). Cumene hydroperoxide
333 (0/8/16/32/64/96/128/160 mM, Sigma-Aldrich, USA) was used as a standard (see more details in
334 Rainio et al. 2019). The absorbance was measured at 570 nm. The results were set against the weight
335 of the body mass of the beetles.

336

337 *2.4. Statistics*

338

339 All statistical analyses were performed with SAS statistical software 9.4 (SAS, 2013) and the figures
340 were prepared with GraphPad Prism 8.4.2. software (GraphPad Prism, 2020). Differences in potato
341 plant glycoalkaloids (α -solanine and α -chaconine) between the treatment groups (GBH-treated and
342 control) were analyzed with repeated generalized linear models (GLMs; Gaussian distribution and
343 identity link function, Glimmix procedure in SAS). Degrees of freedom were calculated with the
344 Kenward-Roger method. The Pearson correlation coefficient was used to test the correlations between
345 potato plant defense chemicals.

346

347 The survival rate of the beetles between the developmental stage (larvae, adults) and treatment groups
348 (GBH-treated, control) and their interaction was analyzed with a generalized linear mixed model
349 (GLMM; with binary distribution and logit link function, events/trials syntax in GLIMMIX

350 procedure, SAS). Family was used as a random factor to control for the non-independence of larvae
351 used from the same family. Degrees of freedom were calculated with the Kenward-Roger method.

352

353 The developmental time of the adult beetles was calculated from hatching of the larvae to newly
354 emerged adult beetles, and the differences in developmental time between the treatment groups was
355 analyzed with a GLMM (Gaussian distribution and identity link function), using treatment (GBH-
356 treated, control), sex (female, male), and treatment \times sex interaction as explanatory variables. Family
357 was used as a random factor. The effect of GBH treatment on body mass (larvae and adults, female
358 and males) was analyzed with a GLMM (Gaussian distribution and identity link function) using
359 family as a random factor.

360

361 To examine the effects of GBH treatment on the oxidative status of the beetles, we performed a
362 GLMM (with lognormal distribution and identity link function, except for CAT and tGSH [for larvae
363 only], in which we used Gaussian distribution and identity link function) for each parameter,
364 separately for larvae and the adult beetles, using treatment (GBH treatment, control), body mass,
365 treatment \times body mass, sex (female, male, adults only), and treatment \times sex (adults only) as
366 explanatory variables. Family was used as a random factor in the models when applicable (larvae:
367 GST, GR, SOD, tGSH, LHP; adults: GP, CAT, tGSH). Non-significant terms were dropped
368 sequentially from the final model, but the main effect of treatment was always kept in the model, as
369 this was our main study question. Degrees of freedom were calculated as mentioned above. Prior to
370 GLMMs, the normality of each parameter was checked. If the parameter was not normally distributed,
371 lognormal distribution was used in the models. The Spearman correlation coefficient was used to test
372 the correlations between oxidative status parameters, body mass, and potato plant glycoalkaloids for
373 larvae and adult beetles, separately in both treatment groups.

374

375 3. Results

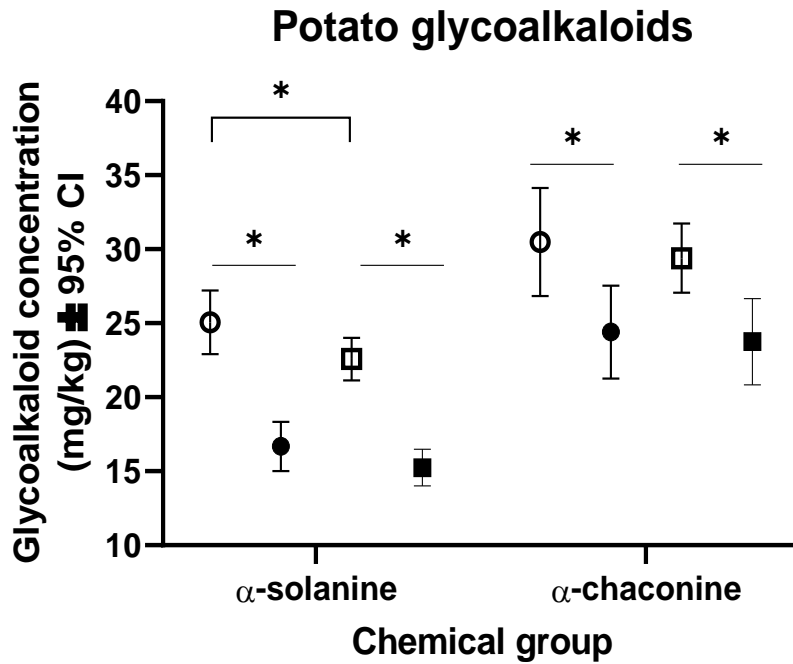
376

377 3.1. Potato plant defense chemicals

378

379 The α -solanine levels were significantly reduced in the potato plants grown in the GBH-treated soil
380 ($F_{df}=6.05_{1, 98}$, $p=0.016$), and the concentrations differed between the measurement times ($F_{df}=98.08_{1, 98}$,
381 $p= <0.001$, Fig. 1), being clearly lower at the second measurement. The treatment \times measurement
382 time interaction was not significant ($F_{df}=0.44_{1, 97}$, $p=0.509$). The α -chaconine levels did not differ
383 between the treatment groups ($F_{df}=0.36_{1, 98}$, $p=0.552$, Fig. 1), but the concentrations differed between
384 the measurement time ($F_{df}=16.17_{1, 98}$, $p=0.0001$, Fig. 1), being likewise lower at the second
385 measurement. There was no significant treatment \times measurement time interaction ($F_{df}=0.02_{1, 97}$,
386 $p=0.880$). The defense chemicals also correlated with each other. The first measurement of α -solanine
387 correlated positively with the second measurement of α -solanine ($r_P^2=0.64$, $p = <0.001$) and with the
388 first measurement of α -chaconine ($r_P^2=0.30$, $p=0.036$); whereas, the second measurement of α -
389 solanine correlated positively with the first ($r_P^2=0.42$, $p=0.002$) and second measurement ($r_P^2=0.74$,
390 $p= <0.001$) of α -chaconine. The first measurement of α -chaconine further correlated positively with
391 the second measurement of α -chaconine ($r_P^2=0.61$, $p= <0.001$).

392



393

394 **Figure 1.** Potato glycoalkaloid (α -solanine and α -chaconine) concentrations (mean \pm 95% CI)
 395 between the treatment groups (GBH treatment, control) at two measurement points (measurement 1,
 396 measurement 2). The color of the symbols indicates measurement time (white=measurement 1,
 397 black=measurement 2) and different symbols the treatment groups (circle=control, square=GBH).
 398 The star above the bars indicate the significant difference between the treatment groups (generalized
 399 linear mixed model, $p < 0.05$).

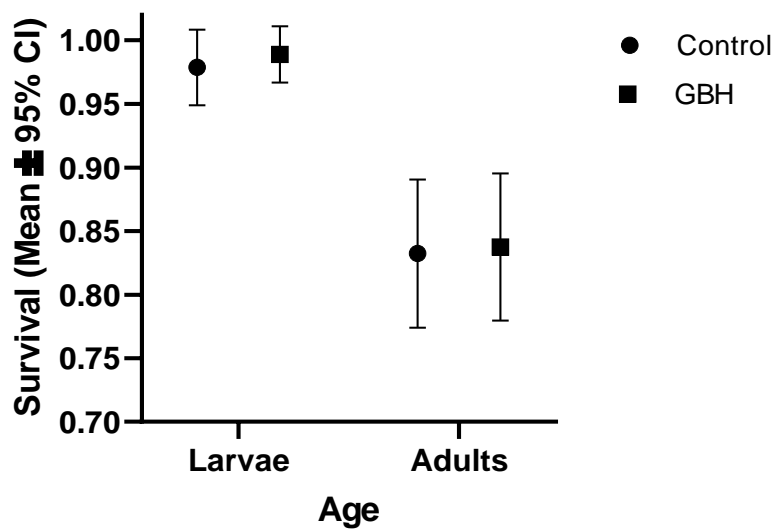
400

401 3.2. Survival rate and changes in developmental time

402

403 GBH treatment had no effect on the survival rate of the Colorado potato beetle larvae or the adult
 404 beetles (Fig. 2). The survival rate of the larvae and the adult beetles differed significantly from each
 405 other, but there was no treatment \times age interaction (Table 1). Larval survival rate in the GBH and
 406 control groups was 98.9% and 97.9%, respectively; whereas, adult survival was 83.9% and 83.4%,
 407 respectively (Table 1). The body mass of the larvae or the adult beetles was not affected by GBH
 408 treatment (larvae: $F_{df}=0.58_{1, 166.2}$, $p=0.447$; adults: $F_{df}=0.01_{1, 254.5}$, $p=0.929$). In the adult beetles,

409 neither the body mass of the females ($F_{df}=0.61_{1, 129.6}$, $p=0.434$) nor males ($F_{df}=0.27_{1, 111.4}$, $p=0.606$)
 410 differed between the treatment groups. However, the developmental time of the adult beetles was
 411 significantly increased in the GBH-treated group compared to the control group (Table 1). Yet, the
 412 estimated difference was only 0.56 days (marginal means: GBH-treated: 30.22, SE: 0.268; control:
 413 29.66, SE: 0.268). Developmental time was not affected by sex or sex \times treatment interaction (Table
 414 1).



415
 416 **Figure 2.** Survival of the Colorado potato beetle (*L. decemlineata*) larvae (2nd instar to 4th instar) and
 417 adults (2nd instar to adult) between the treatment groups (control=black circle, GBH treatment=black
 418 square). The bars represent mean survival (\pm 95% CI) between the treatment groups.

Table 1. The relationship between glyphosate-based herbicide (GBH) treatment and age (larvae and adults) on survival rate of the Colorado potato beetle (<i>L. decemlineata</i>). Significant results are indicated in bold.		
	Survival	
Model*	F_{df}	p
Treatment	0.07 _{1, 502}	0.797
Age	16.93 _{1, 502}	<0.001
Treatment \times age	0.24 _{1, 501}	0.623
	Developmental time	
Model**	F_{df}	p
Treatment	6.26 _{1, 253.2}	0.013
Sex	1.77 _{1, 255.1}	0.185
Treatment \times sex	0.19 _{1, 252.9}	0.667

419 * Generalized linear mixed model (GLMM) with binary distribution and logit link function, family used as a
420 random factor in the model.

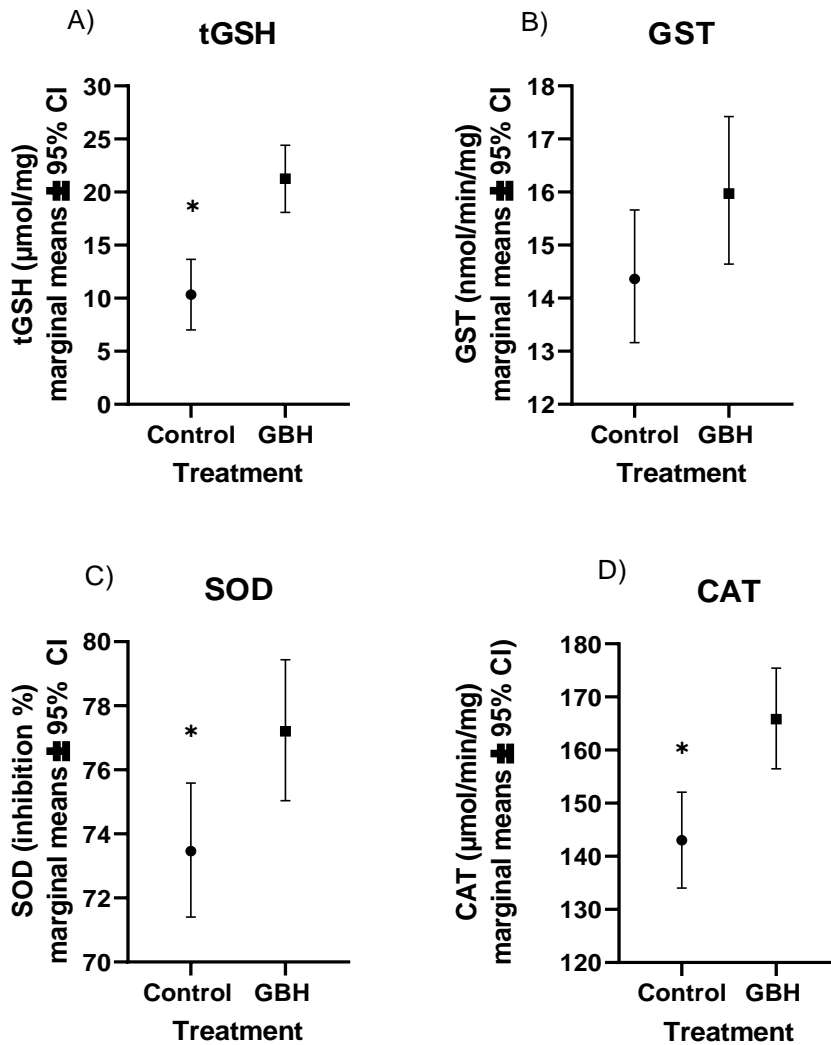
421 ** GLMM with Gaussian distribution and identity link function, family used as a random factor in the
422 model.

423

424 3.3. Oxidative status

425

426 Oxidative status parameters (GR and GPx homologs, GST, tGSH, GSH:GSSG, CAT, SOD and LHP)
427 were analyzed separately between the developmental stages (larvae, adults, Table A1). Oxidative
428 status parameters of the larvae were associated with GBH treatment and body mass, but the body
429 mass \times treatment interaction was not associated with any of the oxidative status parameters (Table
430 2). In the larvae, tGSH concentration and the activity of GST, CAT, and SOD were up-regulated in
431 the GBH treatment group compared to the control group (Table 2, Fig 3.). The other oxidative status
432 parameters (GPx, GR, GSH:GSSG, and LHP) were not associated with GBH treatment. In addition,
433 GST activity was negatively associated with larval body mass, while tGSH concentrations had a
434 positive association with body mass (Table 2). No association between body mass and other oxidative
435 status parameters were found.



436

437 **Figure 3.** Variation in A) total glutathione (tGSH) concentration, B) glutathione-S-transferase (GST),
 438 C) superoxide dismutase (SOD), and D) catalase (CAT) activity in larvae of the Colorado potato
 439 beetle (*L. decemlineata*) between treatment groups (control=black circle, GBH treatment=black
 440 square). The bars represent the marginal means from the models (\pm 95% CI). The star above the bars
 441 indicate significant difference between the treatment groups (generalized linear mixed model,
 442 $p < 0.05$).

443

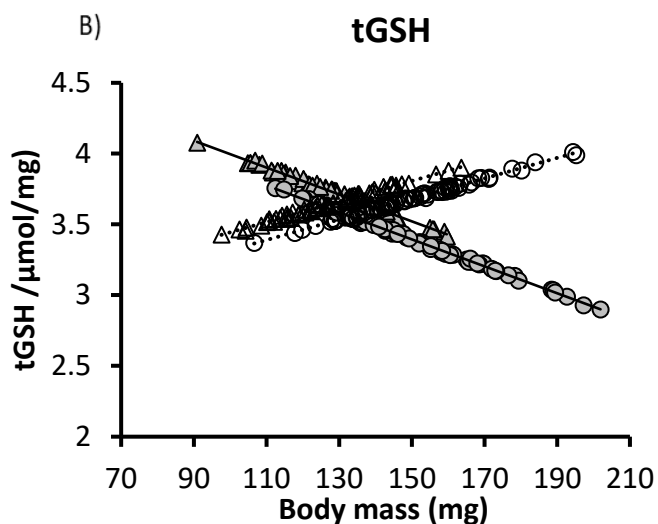
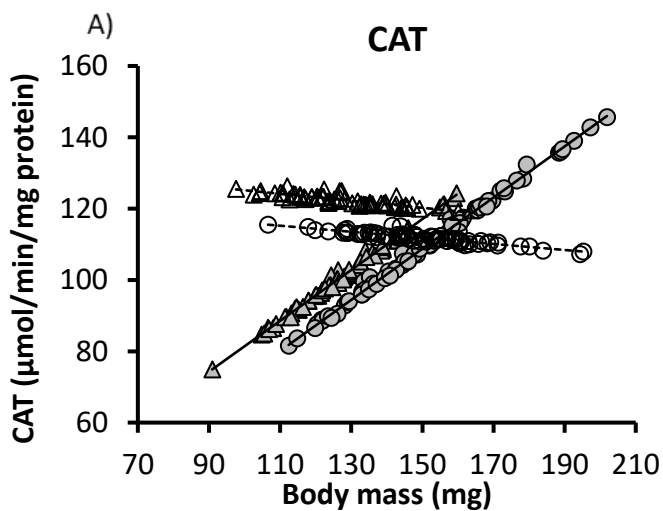
444

445

Table 2. The effects of glyphosate treatment (GBH, control), body mass (bm), sex (female, male), body mass \times treatment, and sex \times treatment interactions on oxidative status parameters glutathione-S-transferase (GST), glutathione peroxidase (GPx), glutathione reductase (GR), catalase (CAT), superoxide dismutase (SOD), total glutathione (tGSH), ratio of reduced and oxidized glutathione (GSH:GSSG) and lipid hydroperoxides (LHP) in larvae and adult Colorado potato beetles (*L. decemlineata*). Non-significant terms were dropped sequentially from each model, starting from interactions (generalized linear mixed model with lognormal distribution and identity link function). Significant results are shown in bold.

Parameters	Model	Larvae			Adults		
		F _{df}	p	n	F _{df}	p	n
GST	treatment	3.88 _{1, 49.97}	0.054	68	0.31 _{1, 60}	0.578	64
	bm	33.99 _{1, 46.49} est. -0.007 SE 0.001	<0.001		4.59 _{1, 60} est. -0.005, SE 0.002	0.036	
	bm*treatment	0.72 _{1, 61.41}	0.399		1.60 _{1, 59}	0.211	
	sex	-	-		1.08 _{1, 60}	0.303	
	sex*treatment	-	-		0.00 _{1, 58}	0.979	
GPx	treatment	0.44 _{1, 65}	0.511	68	0.39 _{1, 43.98}	0.536	61
	bm	0.75 _{1, 65}	0.389		3.48 _{1, 47.6}	0.068	
	bm*treatment	1.02 _{1, 64}	0.316		0.19 _{1, 48.8}	0.669	
	sex	-	-		1.14 _{1, 55.17}	0.289	
	sex*treatment	-	-		0.30 _{1, 43.35}	0.588	
GR	treatment	0.05 _{1, 47.76}	0.823	66	3.39 _{1, 59}	0.071	64
	bm	0.55 _{1, 58.7}	0.460		6.77 _{1, 59} est. 0.003, SE 0.004	0.012	
	bm*treatment	0.47 _{1, 55.41}	0.495		3.33 _{1, 59}	0.073	
	sex	-	-		1.76 _{1, 59}	0.189	
	sex*treatment	-	-		0.04 _{1, 58}	0.842	
CAT	treatment	11.48 _{1, 63}	0.001	65	5.57 _{1, 50.62}	0.022	64
	bm	2.21 _{1, 62}	0.142		1.65 _{1, 48.13}	0.206	
	bm*treatment	1.92 _{1, 61}	0.171		4.61 _{1, 50.81}	0.037	
	sex	-	-		0.81 _{1, 56.95}	0.373	
	sex*treatment	-	-		1.11 _{1, 47.53}	0.297	
SOD	treatment	7.79 _{1, 50}	0.007	68	3.16 _{1, 62}	0.080	64
	bm	0.03 _{1, 46.77}	0.862		1.57 _{1, 61}	0.215	
	bm*treatment	1.80 _{1, 60.44}	0.184		0.28 _{1, 58}	0.599	
	sex	-	-		0.00 _{1, 60}	0.999	
	sex*treatment	-	-		0.43 _{1, 59}	0.512	
tGSH	treatment	42.10 _{1, 32.51}	<.001	43	9.22 _{1, 44.43}	0.004	56
	bm	5.10 _{1, 37.36} est. 0.089, SE 0.039	0.030		0.11 _{1, 42.31}	0.736	
	bm*treatment	1.65 _{1, 38.53}	0.206		10.04 _{1, 44.9}	0.003	
	sex	-	-		0.85 _{1, 48.42}	0.362	
	sex*treatment	-	-		2.68 _{1, 39.84}	0.110	
GSH:GSSG	treatment	1.14 _{1, 41}	0.291	43	0.11 _{1, 51}	0.743	54
	bm	0.15 _{1, 40}	0.704		0.10 _{1, 50}	0.756	
	bm*treatment	0.22 _{1, 39}	0.642		0.38 _{1, 49}	0.543	
	sex	-	-		0.71 _{1, 51}	0.402	
	sex*treatment	-	-		0.00 _{1, 48}	0.991	
LHP	treatment	1.40 _{1, 15.11}	0.255	33	0.01 _{1, 53}	0.908	57
	bm	0.26 _{1, 27.94}	0.613		2.48 _{1, 53}	0.122	
	bm*treatment	0.78 _{1, 18.12}	0.390		0.15 _{1, 52}	0.700	
	sex	-	-		0.57 _{1, 53}	0.452	
	sex*treatment	-	-		0.50 _{1, 51}	0.484	

446 In adult beetles, tGSH concentration and CAT activity had a significant association with treatment \times
 447 body mass interaction (Table 2), and a similar tendency was also found for GR activity (see Table 2).
 448 The GR and CAT activity increased with body mass in the adult beetles in GBH treatment; whereas,
 449 in the control adult beetles, the enzyme activity decreased with increased body mass (Fig. 4). The
 450 tGSH had the opposite trend; the adult beetles in the GBH treatment showed decreased tGSH
 451 concentrations with increased body mass; while in the control, adult beetle tGSH concentrations
 452 increased with body mass (Table 2, Fig. 4). Further, GST activity was negatively associated with
 453 body mass; whereas, GPx had a tendency to be positively associated with body mass (Table 2). No
 454 associations were found for the other measured parameters (SOD, GSH:GSSG, and LHP) of the
 455 oxidative status.



458 **Figure 4.** The relationship between oxidative status parameters (CAT and tGSH) and body mass in
 459 adult Colorado potato beetles (*L. decemlineata*) indirectly exposed to glyphosate (predicted values
 460 from the model; tGSH log transformed values). Legend: white triangle = control male, white
 461 circle = control female, grey triangle = GBH male, grey circle = GBH female.

462

463 We further examined the correlations between the oxidative status parameters and potato plant
 464 glycoalkaloids in the larvae and the adult beetles separately in both treatment groups. We found that
 465 in the GBH treatment group, the CAT activity of the larvae correlated negatively with both
 466 measurements of α -chaconine ($r_s^2=-0.606$, $p=0.028$ and $r_s^2=-0.628$, $p=0.022$, respectively) and with
 467 the second measurement of α -solanine ($r_s^2=-0.694$, $p=0.009$, Table A2 A). Also, the GST levels of
 468 the larvae in the GBH treatment group correlated negatively with the second measurement of α -
 469 solanine and α -chaconine ($r_s^2=-0.558$, $p=0.038$ and $r_s^2=-0.593$, $p=0.025$, respectively, Table A2 A).
 470 There was also a tendency for a negative correlation between GST and the first measurement of α -
 471 chaconine ($r_s^2=-0.513$, $p=0.061$, Table A2 A). The GSH:GSSG ratio had a nearly significant negative
 472 correlation with the first measurement of α -solanine ($r_s^2=-0.592$, $p=0.055$, Table A2 A). The larvae in
 473 the control group had a negative correlation between CAT and the second measurement of α -solanine
 474 ($r_s^2=-0.824$, $p=0.006$), and a nearly significant negative correlation between CAT and the second
 475 measurement of α -chaconine ($r_s^2=-0.656$, $p=0.055$, Table A2 B). There were no significant
 476 correlations between the other parameters ($p>0.05$). In the adult beetles, no correlations between the
 477 potato plant glycoalkaloids and oxidative status parameters were shown in the GBH treatment group
 478 ($p>0.05$, Table A2 C), but in the control group, LHP correlated negatively with the first measurement
 479 of α -solanine ($r_s^2=-0.558$, $p=0.031$, Table A2 D). There were no significant correlations between the
 480 body mass of the larvae and the adult beetles and the potato plant glycoalkaloids ($p>0.05$) in either
 481 of the treatment groups.

482

483 4. Discussion

484

485 4.1. Potato plant defense chemicals

486

487 Soil-mediated exposure to a GBH affected potato plant glycoalkaloid levels. The amount of α -
488 solanine, one of the main defense chemicals of potato plants, was reduced in the potato plants grown
489 in GBH-treated soil compared to the controls; whereas, the α -chaconine levels did not differ between
490 the treatment groups. Correspondingly, Mesnage et al. (2019, preprint) showed in their studies a
491 notable decrease in solanidine (a steroidal alkaloid likewise found in plants of the Solanaceae family)
492 levels in the cecal content of rats exposed to GBH, suggesting that GBH may have a role in the
493 microbial metabolism of alkaloids. GBH has been shown to reduce other secondary compounds in
494 plants as well, such as flavonoid synthesis in barley (*Hordeum vulgare*) seedlings (Laanest, 1987),
495 medicarpin in alfalfa (*Medicago sativa*; Latunde-Dada and Lucas, 1985), and glyceollin in soybeans
496 (*Glycine max*; Ward, 1984). However, opposite results have also been reported, such as the increase
497 of hydrolysable tannins in mountain birch (*Betula pubescens ssp. czerepanovii*; Ossipov et al., 2003).
498 Overall, the effects of GBHs on secondary compounds in plants are surprisingly little studied. The
499 reduction in α -solanine levels may have negative effects on potato plant defense against herbivores,
500 but may benefit the beetles due to lower toxicity of their food items. On the other hand, Colorado
501 potato beetles are specialist herbivores, feeding on *Solanaceae* species with high glycoalkaloid
502 contents, and are well adapted to the defense chemicals of the host plant (Harvey et al., 2005).

503

504 Both α -solanine and α -chaconine levels were reduced in the second measurement compared to the
505 first measurement. The observed difference is most likely related to the size of the potato plant leaves,
506 since the leaves were bigger at the time of the second measurement. Thus, the amount of
507 glycoalkaloids may have become diluted with the leaf growth (personal observations by Rainio and

508 Salminen). However, we cannot entirely rule out the influence of larval feeding or changes caused by
509 potato plant growth on the levels of defense chemicals. For example, Colorado potato beetles have
510 been shown to secrete symbiotic bacteria to suppress plant defenses in tomato plants (*Solanum*
511 *lycopersicum*; Chung et al., 2013), which may apply to potato plant defense chemicals as well.
512 Moreover, GBHs have been shown to affect the growth (Helander et al., 2019) and quality of plants,
513 such as nutrient accumulation (Zobiolo et al., 2012) as well as antioxidant defense (Radwan and
514 Fayed, 2016). For example, glyphosate has been shown to lower photosynthesis and reduce protein-
515 and free amino acid levels as well as induce antioxidant enzyme activities (e.g. CAT, SOD and
516 peroxidases) in peanut (*Arachis hypogaea* L. cv. Giza; Radwan and Fayed, 2016). We did not monitor
517 potato plant growth in this study, but Helander et al. (2019) have shown in their greenhouse
518 experiment that potato plants growing in GBH-treated soil had shorter sprouts soon after planting,
519 but the height of the plants did not differ later during the growing season. However, in the field
520 experiment, the potato plant shoot and tuber biomass was 25% and 14% higher, respectively, from
521 plants grown in GBH-treated soil compared to those grown in control soil (Helander et al., 2019).

522

523 4.2. Survival rate and developmental time

524

525 Soil-mediated exposure to a GBH had no effect on the survival rate of the Colorado potato beetle
526 larvae or the adult beetles, indicating that the environmentally relevant concentrations used in the soil
527 did not increase mortality during the larval stage or show carry-over or soil-mediated effects in adult
528 beetles. The soil used in our experiment contained some glyphosate residues (glyphosate July: 0.41-
529 0.91 mg/kg, AMPA: 0.24-1.00 mg/kg, certified laboratory, Groen Agro Control, Delfgauw,
530 Netherlands, LC-MS/MS, with a detection limit of 0.01 mg/kg). The glyphosate concentrations of the
531 leaves from the present study were not measured, but potato plant leaves, measured from the potato
532 plants grown outside in the field, had no detectable residues (<0.01 mg/kg), unlike potato tubers

533 (glyphosate: 0.02-0.07 mg/kg, AMPA: 0.06-0.07 mg/kg). The adult beetles were also tested for GBH
534 residues to see whether the GBH accumulates in beetles via food at the larval stage or via soil during
535 the pupal phase. Low levels of AMPA were indeed detected in the beetles (AMPA: 0.11mg/kg,
536 glyphosate: 0.013mg/kg), but the residue levels were low and did not affect the survival rate of the
537 beetles at any developmental stage. Our results are in accordance with some other invertebrate
538 studies, which show no effects of GBHs on survival rate (Baker et al., 2014; Haughton et al., 2001;
539 Michalková and Pekár, 2009; Salvio et al., 2016; Thompson et al., 2014). On the other hand, several
540 studies of invertebrates (Benamú et al., 2010; Castilla et al., 2008; Evans et al., 2010; Janssens and
541 Stoks, 2017; Schneider et al., 2009) have shown either direct mortality effects or sublethal effects
542 when exposed to various GBHs, indicating temporal and dose-dependent effects, as well as species-
543 specific differences in insect susceptibility to GBHs. In our earlier study (Rainio et al., 2019), where
544 the Colorado potato beetle larvae were directly exposed to different concentrations of the GBH, low
545 (environmentally relevant) concentrations had no effect on larval survival rate, whereas high
546 concentrations increased larval mortality.

547

548 In the present study, neither the body mass of the larvae or the newly emerged adult beetles (neither
549 females nor males) was affected by GBH treatment, which was expected since the larvae never come
550 in direct contact with the GBH, supporting the finding that the GBH does not affect the beetles'
551 survival rate. However, the developmental time of the adult beetles increased significantly in the
552 GBH treatment group compared to the control group, but the difference (0.56 days) was rather low
553 in a biological sense and likely does not have notable effects on the overall survival rate of the beetles.
554 In general, the Colorado potato beetle tolerates pesticides relatively well, and has developed
555 resistance to several synthetic insecticides, including organophosphates (Kostic et al., 2016; Piironen
556 et al., 2013), used as a control method in potato farms. The metabolic adaptation is manifested by a
557 complex set of detoxifying enzymes, such as GSTs, P450 monooxygenases, and esterases (Ben-

558 Abdallah et al., 2019). Glyphosate also belongs to the organophosphate chemical group, which may
559 potentially affect the susceptibility of the Colorado potato beetles to GBHs. However, this has not
560 been examined in detail.

561

562 4.3. Oxidative status

563

564 Soil-mediated early-life exposure to the GBH affected the antioxidant defense system of the beetles,
565 more specifically the enzymes related to ROS regulation and detoxification of xenobiotics. From the
566 measured oxidative status parameters, GST, CAT, and SOD activity and the concentration of tGSH
567 were up-regulated in the larvae of the GBH-treated group compared to the control group, but this was
568 not seen in the adult stage. The up-regulation can be due to an activation of antioxidant enzymes that
569 work efficiently against increased ROS production to prevent oxidative stress. However, since we did
570 not measure ROS levels, we do not know the exact levels caused by the GBH. On the other hand, it
571 is possible that the potato plant quality (e.g. antioxidant defence, nutrient accumulation) or microbial
572 changes in potato plant (Nissinen et al., unpublished) might have changed due to the GBH treatment,
573 which, in turn, might explain the differences we observe in beetles. In earlier studies, GST activity
574 has been shown to increase in blackworm (*Lumbriculus variegatus*; Contardo-Jara et al., 2009) or
575 decrease in teleostean fish (Samanta et al., 2014) in relation to GBHs or other organophosphorus
576 pesticides e.g. in fish and amphibian studies (Diepens et al., 2014; Oruc, 2011). Insecticide exposure
577 has also been reported to induce GST activity in many insect species (Che-Mendoza et al., 2009).
578 The up-regulation of SOD and CAT activity—the enzymes that catalytically remove ROS (Halliwell
579 and Gutteridge, 2007)—was shown in the larvae, but not in the adult beetles. Since these enzymes
580 operate together, it was expected that they would show a similar trend in relation to GBH treatment.
581 Elevated hepatic SOD and CAT activity has also been found in bullfrog (*Lithobates catesbeiana*)
582 tadpoles exposed to Roundup Original (Costa et al., 2008), increased SOD activity in blackworm

583 exposed to Roundup Ultra (Contardo-Jara et al., 2009), and increased CAT activity in teleost fish
584 exposed to GBHs (Samanta et al., 2014). Our previous direct exposure study of Colorado potato
585 beetles (Rainio et al. 2019) did not show any differences in those same markers of oxidative status,
586 which may be related to the exposure time or the absorption of the GBH by the beetles' bodies
587 (absorption through the cuticle and epidermis vs. via food or soil).

588

589 In addition to enzyme activity, tGSH concentrations in the larvae were elevated in the GBH treatment
590 group compared to the control group. GSH protects cells from oxidative stress by scavenging and
591 neutralizing ROS and simultaneously converting them to GSSG (Halliwell and Gutteridge, 2007;
592 Singh, 2002). The detoxification capacity of GSH is related to its reduced thiol group, and thus the
593 reduced form is the most important in resisting oxidative stress (Singh, 2002). Larsen et al. (2012)
594 reported elevated GSH concentrations in rats exposed to GBHs via drinking water, while some other
595 studies have shown the opposite trend (El-Shenawy, 2009). Increased GSH synthesis, as an adaptive
596 response during moderate oxidative stress, has been previously reported in aquatic organisms by
597 Slaninová et al. (2009). Furthermore, GSH has been suggested to be depleted after short periods of
598 oxidative stress, but elevated after long-term exposure to oxidants (Slaninová et al., 2009). The
599 contradictory results highlight the species- (see also Berglund et al., 2014; Rainio et al., 2013;) and
600 tissue-specificity (Yang et al., 2013) of antioxidant defense, but also the use of various GBHs, the
601 dose and the susceptibility of different species to GBH exposure may induce opposite results. In the
602 present study, the GSH:GSSG ratio and the LHP levels of the larvae did not differ between the
603 treatment groups, suggesting that the increased tGSH level, together with up-regulated enzyme
604 activities, has been effective enough in keeping the cellular redox balance (i.e., GSH:GSSG ratio)
605 stable (Lushchak, 2012). However, the long-term up-regulation of antioxidant enzyme activity is
606 energetically costly and may, in the long-term, increase oxidative stress, ultimately trading-off with
607 the overall survival rate and fitness of the beetles.

608

609 The effect of body mass on oxidative status parameters was further studied in the larvae and the adult
610 beetles, since it has been previously shown that the enzyme activity can be linked to body mass, which
611 is often associated with overall animal condition (Koivula et al., 2011; Rainio et al., 2015). In the
612 larvae (as also in the adults), the body mass had a negative association with GSTs, meaning that the
613 lighter larvae had higher GST activity compared to heavier larvae. It is possible that, in general, the
614 lighter larvae that are in poorer condition need to up-regulate GST activity more for detoxification
615 processes, which may be energy demanding, than the heavier ones that are in better condition. A
616 similar results between the antioxidant enzyme activities of GPx, SOD, and CAT and body mass have
617 been found in birds, such as the great tit (*Parus major*), when exposed to metal pollution (Rainio et
618 al., 2015). The larvae further showed a positive association between body mass and tGSH
619 concentrations, meaning that heavier larvae had higher tGSH levels, which is opposite to what we
620 found for GST. However, it may be that the heavier larvae can produce more GSH in their system,
621 reflecting better antioxidant capacity, compared to the lighter larvae that are in poorer condition.

622

623 In this study, we were able to follow the individuals from the larvae to the adult stage to examine the
624 long-term effects of early-life GBH exposure. The GBH directly decreased the oxidative status
625 parameters CAT and tGSH in the adult beetles, and there was a significant treatment \times body mass
626 interaction. In the adult beetles, CAT activity (and GR activity to some extent) increased with body
627 mass in the GBH treatment group, but decreased in the control group. The opposite was shown for
628 tGSH, where the levels increased with body mass in the control group, but decreased in the GBH
629 treatment group. The higher CAT activity of the heavier adult beetles in the GBH treatment group
630 may be due to being in better condition, allowing them to allocate more resources for their defense in
631 case of increased ROS production compared to lighter ones that are in poorer condition. However, in
632 the controls, the body mass may not be so critical since their activities stay rather constant.

633

634 The increased tGSH levels may reflect the better condition of heavier adults in the control group;
635 whereas, in the GBH treatment group, the decreased tGSH levels may suggest either lesser need of
636 tGSH (e.g. due to up-regulated enzyme activities) or more rapid transformation of GSH to GSSG to
637 cope with the potential increase in ROS production. This is further supported by the higher GR
638 activity in the bigger adults than the smaller ones in the GBH treatment group, since the main function
639 of GR is to transform oxidized GSH (i.e. GSSG) back to its reduced form (GSH; Halliwell and
640 Gutteridge, 2007). The results suggest that the early-life indirect GBH exposure via diet may show
641 some long-term effects on the adult beetles. On the other hand, the pupa may also be directly exposed
642 to GBH residues during their 2-week pupal stage in the soil, which can partly explain the observed
643 effects on the adults' physiology and developmental time between the treatment groups. In future, it
644 would be important to concentrate more on the plant-mediated effects and separate them from the
645 soil-mediated effects at the pupal stage, and, moreover, extend the studies to observe the following
646 breeding season to see whether the GBH affects the overwintering and reproduction success of the
647 adult beetles later in life.

648

649 We also examined the relationships between oxidative status parameters and potato plant
650 glycoalkaloids separately in larvae and the adult beetles to see whether these chemicals affect the
651 beetle's oxidative status. We found that for the larvae in the GBH treatment group (as also in the
652 control group), the activity of CAT and GST correlated negatively with α -solanine and α -chaconine
653 levels, either with both of the measurements (before and after larval feeding) or with only one of the
654 measurements. Interestingly, these are the same parameters that were affected by GBH treatment in
655 larvae, but in the opposite direction. The GST and CAT activity decreased with increased α -solanine
656 and α -chaconine levels, but increased with GBH treatment. The results are logical, since the lower α -
657 solanine levels were shown in the GBH treatment group with higher antioxidant enzyme activity. The

658 observed changes in antioxidant defense of the beetles can be derived from the GBH itself or from
659 the GBH-mediated effects on potato glycoalkaloid levels, in case the glycoalkaloids affect the potato
660 quality as food items. The α -solanine has been previously shown to increase lipid peroxidation
661 (measured as malondialdehyde [MDA] concentration) and GST activity in the mid-gut, but decrease
662 the GST activity in body fat in Lepidoptera, such as *G. mellonella*, indicating the oxidative activity
663 of glycoalkaloids (Adamski et al., 2014). Furthermore, GSH:GSSG ratio had a similar tendency for
664 a negative correlation with only the first measurement of α -solanine (see table S2), reflecting the
665 increased oxidation of GSH to GSSG in the higher concentrations of glycoalkaloids. In the adult
666 beetles, on the other hand, none of the oxidative status parameters correlated with potato plant
667 glycoalkaloids. Even though both potato plant defense chemicals and GBH treatment seemed to affect
668 the same oxidative status parameters of the beetle larvae (e.g., GST, CAT), we cannot say for sure
669 whether they show additive or synergistic effects on the beetles. More experimental studies with
670 different concentrations of glycoalkaloids and GBHs would be needed to understand the complex
671 combined effects of glycoalkaloids and GBHs on the oxidative status parameters of the beetles.

672

673 4.4. Conclusions

674

675 The reduction of α -solanine levels in potato plants grown in GBH-treated soil suggests the potential
676 reduction of potato plant defense against the Colorado potato beetle, but more dose-dependent studies
677 would be needed to examine the significance of the reduction of defense chemicals on potato plants,
678 since the herbicides may significantly affect the inter- and intraspecies interactions of agricultural
679 ecosystems. The survival rate of the beetles was not affected by the soil-mediated early-life GBH
680 treatment, but the oxidative status parameters, GST, SOD, CAT, and tGSH, were increased in the
681 larvae in the GBH treatment group compared to the control group. The long-term up-regulation of
682 antioxidant enzyme activity is energetically costly and may increase oxidative stress in the larvae,

683 which could in turn delay the developmental time. In the adult beetles, CAT activity and tGSH levels
684 were affected by the interactive effect of GBH treatment and body mass of the adult beetles,
685 suggesting that the early-life glyphosate treatment or soil-mediated effects at the pupal stage may
686 have long-term effects on the adult beetles. Our results highlight the importance of measuring the
687 physiological parameters, such as oxidative status, along with life-history traits in sublethal herbicide
688 studies, since they may be important factors in affecting the health and survival of animals. In future,
689 it would be important to extend the monitoring of the adult beetles to the following breeding season,
690 to study the effects of GBHs on fertility, reproductive success, and overwinter survival rate of the
691 adult beetles.

692

693 **Acknowledgements**

694

695 We would like to thank Maija Jortikka, Anna Pauna, and Otto Saikkonen for their help in rearing the
696 beetles. This study was funded by the Academy of Finland (grant no. 311077 to MH), the Alfred
697 Kordelin Foundation (MR), and the Tiina and Antti Herlin Foundation (MR).

698

699 **Conflicts of Interest**

700

701 The authors declare no conflict of interest.

702

703 **Credit Author Statement**

704

705 **Miia J. Rainio:** Study design, conducting experiment, biochemical analyses, statistical analyses,
706 manuscript writing. **Aigi Margus:** Study design, experiment preparation, manuscript editing.

707 **Valtteri Virtanen:** Glycoalkaloid analyses, manuscript editing. **Leena Lindström:** Study design,

708 experiment preparation, manuscript editing. **J-P Salminen:** Glycoalkaloid analyses, manuscript
 709 editing. **Kari Saikkonen:** manuscript editing. **Marjo Helander:** Study design, manuscript editing.

710

711 **References**

712 Adamski, Z., Marciniak, P., Ziemnicki, K., Büyükgüzel, E., Erdem, M., Büyükgüzel, K., Ventrella,
 713 E., Falabella, P., Cristallo, M., Salvia, R., Bufo, S.A., Scrano, L., 2014. Potato leaf extract and its
 714 component, α -solanine, exert similar impacts on development and oxidative stress in *Galleria*
 715 *mellonella* L. *Archives of Insect Biochemistry and Physiology* 87(1), 26-39.

716

717 Alghamdi, A.A., Frey, K.M., 2017. Predicting The Toxic Effect of Organophosphates on GST
 718 Enzyme Isoforms. *The FASEB Journal* 31, 1b623-1b623.

719

720 Amrhein, N., Deus, B., Gehrke, P., Steinrücken, H.C., 1980. The site of the inhibition of the
 721 shikimate pathway by glyphosate: II. Interference of glyphosate with chorismate formation in vivo
 722 and in vitro. *Plant Physiology* 66(5), 830-834.

723

724 Andrews, G.K., 2000. Regulation of metallothionein gene expression by oxidative stress and metal
 725 ions. *Biochemical Pharmacology* 59(1), 95-104.

726

727 Annett, R., Habibi, H.R., Hontela, A., 2014. Impact of glyphosate and glyphosate-based herbicides
 728 on the freshwater environment. *Journal of Applied Toxicology* 34(5), 458-479.

729

730 Baker, L.F., Mudge, J.F., Houlahan, J.E., Thompson, D.G., Kidd, K.A., 2014. The direct and
 731 indirect effects of a glyphosate-based herbicide and nutrients on *Chironomidae* (Diptera) emerging
 732 from small wetlands. *Environmental Toxicology and Chemistry* 33, 2076-2085.

733

734 Ben-Abdallah, S., Cáceres, L.A., Wang, Z.L., Renaud, B.J., Lachâal, M., Karray-Bouraoui, N.,
 735 Hannoufa, A., Scott, I.M., 2019. Host plant defenses of black (*Solanum nigrum* L.) and red
 736 nightshade (*Solanum villosum* Mill.) against specialist Solanaceae herbivore *Leptinotarsa*
 737 *decemlineata* (Say). *Archives of Insect Biochemistry and Physiology* 101(2).

738

739 Benamú, M.A., Schneider, M.I., Sánchez, N.E., 2010. Effects of the herbicide glyphosate on
 740 biological attributes of *Alpaida veniliae* (Araneae, Araneidae), in laboratory. *Chemosphere* 78(7),
 741 871-876.

742

743 Bentley, R., 1990. The shikimate pathway - a metabolic tree with many branches. *Critical Reviews*
 744 *in Biochemistry and Molecular Biology* 25(5), 307-384.

745

746 Berglund, Å.M.M., Rainio, M.J., Kanerva, M., Nikinmaa, M., Eeva, T., 2014. Antioxidant status in
 747 relation to age, condition, reproductive performance and pollution in three passerine species.
 748 *Journal of Avian Biology* 45(3), 235-246.

749

750 Blumenthal, D., Augustine, D., 2009. Plant interactions with herbivores. *Encyclopedia of the Life*
 751 *Sciences*.

752

- 753 Bou, R., Codony, R., Tres, A., Decker, E.A., Guardicila, F., 2008. Determination of hydroperoxides
754 in foods and biological samples by the ferrous oxidation-xylene orange method: A review of the
755 factors that influence the method's performance. *Analytical Biochemistry* 377(1), 1-15.
756
- 757 Casagrande, R.A., 1987. The Colorado potato beetle: 125 years of mismanagement. *Bulletin of the*
758 *Entomological Society of America* 33(3), 142-150.
759
- 760 Castilla, A.M., Dauwe, T., Mora, I., Malone, J., Guitart, R., 2010. Nitrates and herbicides cause
761 higher mortality than the traditional organic fertilizers on the grain beetle, *Tenebrio molitor*.
762 *Bulletin of Environmental Contamination and Toxicology* 84(1), 101-105.
763
- 764 Castilla, A.M., Dauwe, T., Mora, I., Palmer, M., Guitart, R., 2008. Mortality of the yellow
765 mealworm *Tenebrio molitor* exposed to fertilizers and herbicides commonly used in agriculture.
766 *Vie Et Milieu-Life and Environment* 58(3-4), 243-247.
767
- 768 Che-Mendoza, A., Penilla, R., Rodríguez, D., 2009. Insecticide resistance and glutathione S-
769 transferases in mosquitoes: A review. *African Journal of Biotechnology* 8(8), 1386-1397.
770
- 771 Chowański, S., Adamski, Z., Marciniak, P., Rosiński, G., Büyükgüzel, E., Büyükgüzel, K.,
772 Falabella, P., Scrano, L., Ventrella, E., Lelario, F., Bufo, S., 2016. A review of bioinsecticidal
773 activity of Solanaceae alkaloids. *Toxins (Basel)* 8(3), 60.
774
- 775 Chung, S.H., Rosa, C., Hoover, K., Luthe, D.S., Felton, G.W., 2013. Colorado potato beetle
776 manipulates plant defenses in local and systemic leaves. *Plant Signaling & Behavior* 8(12), e27592.
777
- 778 Claus, S.P., Guillou, H., Ellero-Simatos, S., 2016. The gut microbiota: a major player in the toxicity
779 of environmental pollutants? *npj Biofilms and Microbiomes* 2, 16003.
780
- 781 Contardo-Jara, V., Klingelmann, E., Wiegand, C., 2009. Bioaccumulation of glyphosate and its
782 formulation Roundup Ultra in *Lumbriculus variegatus* and its effects on biotransformation and
783 antioxidant enzymes. *Environmental Pollution* 157(1), 57-63.
784
- 785 Corona, M., Robinson, G.E., 2006. Genes of the antioxidant system of the honey bee: annotation
786 and phylogeny. *Insect Molecular Biology* 15(5), 687-701.
787
- 788 Costa, M.J., Monteiro, D.A., Oliveira-Neto, A.L., Rantin, F.T., Kalinin, A.L., 2008. Oxidative
789 stress biomarkers and heart function in bullfrog tadpoles exposed to Roundup Original®.
790 *Ecotoxicology* 17(3), 153-163.
791
- 792 Deisseroth, A., Dounce, A.L., 1970. Catalase: physical and chemical properties, mechanism of
793 catalysis, and physiological role. *Physiological Reviews* 50(3), 319-375.
794
- 795 Dewick, P., 2009. *Medicinal Natural Products: A Biosynthetic Approach*, 3rd ed. John Wiley and
796 Sons Ltd, United Kingdom.
797
- 798 Diepens, N., Pfennig, S., Van den Brink, P., Gunnarsson, J., Ruepert, C., Castillo, L., 2014. Effect
799 of pesticides used in banana and pineapple plantations on aquatic ecosystems in Costa Rica. *Journal*
800 *of Environmental Biology* 35(1), 73-84.
801

- 802 Downer, R.G.H., 1985. Lipid metabolism, in: Kerkut, G.A., Gilbert L. I. (Eds.), Comprehensive
803 insect physiology, biochemistry and pharmacology, Pergamon Press, Oxford, United Kingdom, pp.
804 77-113.
- 805
- 806 Duke, S.O., Powles, S.B., 2008. Glyphosate: A once-in-a-century herbicide. *Pest Management*
807 *Science* 64(4), 319-325.
- 808
- 809 El-keltawi, N.E., Croteau, R., 1987. Influence of foliar applied cytokinins on growth and essential
810 oil content of several members of the lamiaceae. *Phytochemistry* 26(4), 891-895.
- 811
- 812 El-Shenawy, N.S., 2009. Oxidative stress responses of rats exposed to Roundup and its active
813 ingredient glyphosate. *Environmental Toxicology and Pharmacology* 28(3), 379-385.
- 814
- 815 Evans, S.C., Shaw, E.M., Rypstra, A.L., 2010. Exposure to a glyphosate-based herbicide affects
816 agrobiont predatory arthropod behaviour and long-term survival. *Ecotoxicology* 19(7), 1249-1257.
- 817
- 818 Farjan, M., Dmitryjuk, M., Lipinski, Z., Biernat-Lopienska, E., Zoltowska, K., 2012.
819 Supplementation of the honey bee diet with vitamin C: The effect on the antioxidative system of
820 *Apis mellifera carnica* brood at different stages. *Journal of Apicultural Research* 51(3), 263-270.
- 821
- 822 Finkel, T., Holbrook, N.J., 2000. Oxidants, oxidative stress and the biology of ageing. *Nature*
823 408(6809), 239-247.
- 824
- 825 Fossati, P., Prencipe, L., Berti, G., 1980. Use of 3,5-dichloro-2-hydroxybenzenesulfonic acid/4-
826 aminophenazone chromogenic system in direct enzymic assay of uric-acid in serum and urine.
827 *Clinical Chemistry* 26(2), 227-231.
- 828
- 829 Fragoyiannis, D., McKinlay, R., D'Mello, J., 1998. Studies of the growth, development and
830 reproductive performance of the aphid shape *Myzus persicae* on artificial diets containing potato
831 glycoalkaloids. *Entomologia Experimentalis et Applicata* 88, 59-66.
- 832
- 833 Friedman, M., 2006. Potato glycoalkaloids and metabolites: roles in the plant and in the diet.
834 *Journal of Agricultural and Food Chemistry* 54(23), 8655-8681.
- 835
- 836 Friedman, M., McDonald, G.M., Filadelfi-Keszi, M., 1997. Potato Glycoalkaloids: Chemistry,
837 Analysis, Safety, and Plant Physiology. *Critical Reviews in Plant Sciences* 16(1), 55-132.
- 838
- 839 Funke, T., Han, H., Healy-Fried, M.L., Fischer, M., Schönbrunn, E., 2006. Molecular basis for the
840 herbicide resistance of Roundup Ready crops. *Proceedings of the National Academy of Sciences of*
841 *the United States of America* 103(35), 13010-13015.
- 842
- 843 George, D.G.M., Gatehouse, A.M.R., 2013. Oxidative stress enzymes in *Busseola fusca*.
844 *International Journal of Current Microbiology and Applied Sciences* 2(10), 485-495.
- 845
- 846 Giesy, J.P., Dobson, S., Solomon, K.R., 2000. Ecotoxicological risk assessment for Roundup (R)
847 Herbicide. In: Ware, G.W. (Ed.). *Reviews of Environmental Contamination and Toxicology*, Vol
848 167, Springer, New York, 35-120.
- 849
- 850 Glass, R.L., 1984. Metal complex formation by glyphosate. *Journal of Agricultural and Food*
851 *Chemistry* 32(6), 1249-1253.

- 852
853 Glusczak, L., Miron, D.D., Moraes, B.S., Simoes, R.R., Schetinger, M.R.C., Morsch, V.M., Loro,
854 V.L., 2007. Acute effects of glyphosate herbicide on metabolic and enzymatic parameters of silver
855 catfish (*Rhamdia quelen*). *Comparative Biochemistry and Physiology - Part C: Toxicology &*
856 *Pharmacology* 146(4), 519-524.
857
- 858 Gomes, M.P., Le Manac'h, S.G., Maccario, S., Labrecque, M., Lucotte, M., Juneau, P., 2016.
859 Differential effects of glyphosate and aminomethylphosphonic acid (AMPA) on photosynthesis and
860 chlorophyll metabolism in willow plants. *Pesticide Biochemistry and Physiology* 130, 65-70.
861
- 862 GraphPad Prism, 2020. User Guide. GraphPad Software, LLC.
863
- 864 Grapputo, A., Boman, S., Lindstrom, L., Lyytinen, A., Mappes, J., 2005. The voyage of an invasive
865 species across continents: genetic diversity of North American and European Colorado potato beetle
866 populations. *Molecular Ecology* 14(14), 4207-4219.
867
- 868 Güntner, C., Gonzalez, A., Dos Reis, R., Usubillanga, A., Ferreira, F., Moyna, P., 1997. Effect of
869 Solanum glycoalkaloids on potato aphid, *Macrosiphum euphorbiae*. *Journal of Chemical Ecology*
870 23, 1651-1659.
871
- 872 Hagner, M., Mikola, J., Saloniemi, I., Saikkonen, K., Helander, M., 2019. Effects of a glyphosate-
873 based herbicide on soil animal trophic groups and associated ecosystem functioning in a northern
874 agricultural field. *Scientific Reports* 9, 8540.
875
- 876 Halliwell, B., Gutteridge, J., 2007. Free Radicals in Biology and Medicine, Fourth ed. Oxford
877 University Press, New York.
878
- 879 Hare, J.D., 1987. Growth of *Leptinotarsa decemlineata* larvae in response to simultaneous variation
880 in protein and glycoalkaloid concentration. *Journal of Chemical Ecology* 13, 39-46.
881
- 882 Harvey, J.A., van Nouhuys, S., Biere, A., 2005. Effects of quantitative variation in allelochemicals
883 in *Plantago lanceolata* on development of a generalist and a specialist herbivore and their
884 endoparasitoids. *Journal of Chemical Ecology* 31(2), 287-302.
885
- 886 Haslam, E., 1993. Shicimic Acid: Metabolism and Metabolites, 1 ed. Wiley, New York.
887
- 888 Haughton, A.J., Bell, J.R., Wilcox, A., Boatman, N.D., 2001. The effect of the herbicide glyphosate
889 on non-target spiders: Part I. Direct effects on *Lepthyphantes tenuis* under laboratory conditions.
890 *Pest Management Science* 57(11), 1033-1036.
891
- 892 Helander, M., Pauna, A., Saikkonen, K., Saloniemi, I., 2019. Glyphosate residues in soil affect crop
893 plant germination and growth. *Scientific Reports* 9, 19653.
894
- 895 Helander, M., Saloniemi, I., Omacini, M., Druille, M., Salminen, J.-P., Saikkonen, K., 2018.
896 Glyphosate decreases mycorrhizal colonization and affects plant-soil feedback. *Science of the Total*
897 *Environment* 642, 285-291.
898
- 899 Helander, M., Saloniemi, I., Saikkonen, K., 2012. Glyphosate in northern ecosystems. *Trends in*
900 *Plant Science* 17(10), 569-574.
901

- 902 Hultberg, M., 2007. Cysteine turnover in human cell lines is influenced by glyphosate.
903 *Environmental Toxicology and Pharmacology* 24(1), 19-22.
904
- 905 Isaksson, C., Oernborg, J., Stephensen, E., Andersson, S., 2005. Plasma glutathione and carotenoid
906 coloration as potential biomarkers of environmental stress in great tits. *EcoHealth* 2(2), 138-146.
907
- 908 Janssens, L., Stoks, R., 2017. Stronger effects of Roundup than its active ingredient glyphosate in
909 damselfly larvae. *Aquatic Toxicology* 193, 210-216.
910
- 911 Khan, M., Munir, I., Khan, I., 2013. The potential of unintended effects in potato glycoalkaloids.
912 *African*
913 *Journal of Biotechnology* 12(8), 754-766.
914
- 915 Kishore, G.M., Shah, D.M., 1988. Amino-acid biosynthesis inhibitors as herbicides. *Annual Review*
916 *of Biochemistry* 57, 627-663.
917
- 918 Koivula, M.J., Kanerva, M., Salminen, J.P., Nikinmaa, M., Eeva, T., 2011. Metal pollution
919 indirectly increases oxidative stress in great tit (*Parus major*) nestlings. *Environmental Research*
920 111(3), 362-370.
921
- 922 Kostic, M., Stankovic, S., Kuzevski, J., 2016. Role of AChE in Colorado potato beetle
923 (*Leptinotarsa decemlineata* Say) Resistance to Carbamates and Organophosphates. InTechOpen.
924 Retrieved from [https://www.intechopen.com/books/insecticides-resistance/role-of-ache-in-](https://www.intechopen.com/books/insecticides-resistance/role-of-ache-in-colorado-potato-beetle-leptinotarsa-decemlineata-say-resistance-to-carbamates-and-or)
925 [colorado-potato-beetle-leptinotarsa-decemlineata-say-resistance-to-carbamates-and-or](https://www.intechopen.com/books/insecticides-resistance/role-of-ache-in-colorado-potato-beetle-leptinotarsa-decemlineata-say-resistance-to-carbamates-and-or)
926
- 927 Kowalski, S.P., Domek, J.M., Deahl, K.L., Sanford, L.L., 1999. Performance of Colorado potato
928 beetle larvae, *Leptinotarsa decemlineata* (Say), reared on synthetic diets supplemented with
929 *Solanum* glycoalkaloids. *American Journal of Potato Research* 76, 305-312.
930
- 931 Laanest, L., 1987. Incorporation of exogenous tyrosine and phenylalanine into C-glycosylflavones
932 in glyphosate-treated barley seedlings. *Eesti NSV Teaduste Akadeemia Toimetised Bioloogia* 36(3),
933 204-209.
934
- 935 Lachman, J., Hamouz, K., Orsak, M., Pivec, V., 2001. Potato glycoalkaloids and their significance
936 in plant protection and human nutrition - review. *Rostlinna Výroba* 47(4), 181-191.
937
- 938 Laitinen, P., 2009. Glyphosate and phosphorus leaching and residues in boreal sandy soil. *Plant and*
939 *Soil* 323(1), 267-283.
940
- 941 Larsen, K., Najle, R., Lifschitz, A., Virkel, G., 2012. Effects of sub-lethal exposure of rats to the
942 herbicide glyphosate in drinking water: Glutathione transferase enzyme activities, levels of reduced
943 glutathione and lipid peroxidation in liver, kidneys and small intestine. *Environmental Toxicology*
944 *and Pharmacology* 34(3), 811-818.
945
- 946 Latunde-Dada, A.O., Lucas, J.A., 1985. Involvement of the phytoalexin medicarpin in the
947 differential response of callus lines of lucerne (*Medicago sativa*) to infection by *Verticillium albo-*
948 *atrum*. *Physiological Plant Pathology* 26(1), 31-42.
949

- 950 Lehmann, P., Lyytinen, A., Piironen, S., Lindstrom, L., 2015. Latitudinal differences in diapause
951 related photoperiodic responses of European Colorado potato beetles (*Leptinotarsa decemlineata*).
952 *Evolutionary Ecology* 29(2), 269-282.
953
- 954 Liu, X., Williams, C.E., Nemacheck, J.A., Wang, H., Subramanyam, S., Zheng, C., Chen, M.-S.,
955 2010. Reactive oxygen species are involved in plant defense against a gall midge. *Plant Physiology*
956 152(2), 985-999.
957
- 958 Lushchak, V.I., 2012. Glutathione homeostasis and functions: potential targets for medical
959 interventions. *Journal of Amino Acids* 2012, 1-26.
960
- 961 Lydon, J., Duke, S.O., 1989. Pesticide effects on secondary metabolism of higher-plants. *Pesticide*
962 *Science* 25(4), 361-373.
963
- 964 Lyytinen, A., Lindstrom, L., Mappes, J., Julkunen-Tiitto, R., Fasulati, S.R., Tiilikkala, K., 2007.
965 Variability in host plant chemistry: behavioural responses and life-history parameters of the
966 Colorado potato beetle (*Leptinotarsa decemlineata*). *Chemoecology* 17, 51-56.
967
- 968 Margus, A., Rainio, M., Lindström, L., 2019. Can indirect herbicide exposure modify the response
969 of the Colorado potato beetle to an organophosphate insecticide? *Journal of Economic Entomology*
970 112(5), 2316-2323.
971
- 972 Martinez, D.A., Loening, U.E., Graham, M.C., 2018. Impacts of glyphosate-based herbicides on
973 disease resistance and health of crops: a review. *Environmental Sciences Europe* 30(1), 2.
974
- 975 Matthews, D., Jones, H., Gans, P., Coates, S., Smith, L.M.J., 2005. Toxic secondary metabolite
976 production in genetically modified potatoes in response to stress. *Journal of Agricultural and Food*
977 *Chemistry* 53(20), 7766-7776.
978
- 979 Mesnage, R., Teixeira, M., Madrioli, D., Falcioni, L., Ducarmon, Q.R., Zwittink, R.D., Amiel, C.,
980 Panoff, J.-M., Belpoggi, F., Antoniou, M.N. 2019. Shotgun metagenomics and metabolomics reveal
981 glyphosate alters the gut microbiome of Sprague-Dawley rats by inhibiting the shikimate pathway.
982 *BioRxiv* preprint. doi: <https://doi.org/10.1101/870105>.
983
- 984 Mesnage, R., Defarge, N., de Vendômois, J.S., Séralini, G.E., 2015. Potential toxic effects of
985 glyphosate and its commercial formulations below regulatory limits. *Food and Chemical*
986 *Toxicology* 84, 133-153.
987
- 988 Mesnage, R., Defarge, N., Spiroux de Vendômois, J., Séralini, G.-E., 2014. Major pesticides are
989 more toxic to human cells than their declared active principles. *BioMed Research International*
990 2014, 179691.
991
- 992 Michalková, V., Pekár, S., 2009. How glyphosate altered the behaviour of agrobiont spiders
993 (Araneae: Lycosidae) and beetles (Coleoptera: Carabidae). *Biological Control* 51(3), 444-449.
994
- 995 Modesto, K.A., Martinez, C.B.R., 2010. Roundup causes oxidative stress in liver and inhibits
996 acetylcholinesterase in muscle and brain of the fish *Prochilodus lineatus*. *Chemosphere* 78(3), 294-
997 299.
998

- 999 Myers, J.P., Antoniou, M.N., Blumberg, B., Carroll, L., Colborn, T., Everett, L.G., Hansen, M.,
1000 Landrigan, P.J., Lanphear, B.P., Mesnage, R., Vandenberg, L.N., vom Saal, F.S., Welshons, W.V.,
1001 Benbrook, C.M., 2016. Concerns over use of glyphosate-based herbicides and risks associated with
1002 exposures: a consensus statement. *Environmental Health* 15, 19.
1003
- 1004 Nenaah, G., 2011. Toxic and antifeedant activities of potato glycoalkaloids against *Trogoderma*
1005 *granarium* (Coleoptera: Dermestidae). *Journal of Stored Products Research* 47(3), 185-190.
1006
- 1007 Nourooz-Zadeh, J., Tajaddini-Sarmadi, J., McCarthy, S., Betteridge, D.J., Wolff, S.P., 1995.
1008 Elevated levels of authentic plasma hydroperoxides in NIDDM. *Diabetes* 44(9), 1054.
1009
- 1010 Nylin, S., Janz, N., 1993. Oviposition preference and larval performance in *Polygona c-album*
1011 (Lepidoptera: Nymphalidae): the choice between bad and worse. *Ecological Entomology* 18(4),
1012 394-398.
1013
- 1014 Oruc, E., 2011. Effects of diazinon on antioxidant defense system and lipid peroxidation in the liver
1015 of *Cyprinus carpio* (L.). *Environmental Toxicology* 26(6), 571-578.
1016
- 1017 Ossipov, V., Salminen, J.-P., Ossipova, S., Haukioja, E., Pihlaja, K., 2003. Gallic acid and
1018 hydrolysable tannins are formed in birch leaves from an intermediate compound of the shikimate
1019 pathway. *Biochemical Systematics and Ecology* 31(1), 3-16.
1020
- 1021 Piironen, S., Lindstrom, L., Lyytinen, A., Mappes, J., Chen, Y.H., Izzo, V., Grapputo, A., 2013.
1022 Pre-invasion history and demography shape the genetic variation in the insecticide resistance-
1023 related acetylcholinesterase 2 gene in the invasive Colorado potato beetle. *BMC Evolutionary*
1024 *Biology* 13, 13.
1025
- 1026 Pinto, E., Sigaud-Kutner, T.C.S., Leitao, M.A.S., Okamoto, O.K., Morse, D., Colepicolo, P., 2003.
1027 Heavy metal-induced oxidative stress in algae. *Journal of Phycology* 39(6), 1008-1018.
1028
- 1029 Radwan, D.E.M., Fayez, K.A., 2016. Photosynthesis, antioxidant status and gas-exchange are
1030 altered by glyphosate application in peanut leaves. *Photosynthetica* 54, 307-316.
1031
- 1032 Rainio, M.J., Eeva, T., Lilley, T., Stauffer, J., Ruuskanen, S., 2015. Effects of early-life lead
1033 exposure on oxidative status and phagocytosis activity in great tits (*Parus major*). *Comparative*
1034 *Biochemistry and Physiology - Part C: Toxicology & Pharmacology* 167, 24-34.
1035
- 1036 Rainio, M.J., Kanerva, M., Salminen, J.-P., Nikinmaa, M., Eeva, T., 2013. Oxidative status in
1037 nestlings of three small passerine species exposed to metal pollution. *Science of the Total*
1038 *Environment* 454-455, 466-473.
1039
- 1040 Rainio, M.J., Margus, A., Lehmann, P., Helander, M., Lindström, L., 2019. Effects of a glyphosate-
1041 based herbicide on survival and oxidative status of a non-target herbivore, the Colorado potato
1042 beetle (*Leptinotarsa decemlineata*). *Comparative Biochemistry and Physiology - Part C:*
1043 *Toxicology & Pharmacology* 215, 47-55.
1044
- 1045 Salvio, C., Menone, M.L., Rafael, S., Iturburu, F.G., Manetti, P.L., 2016. Survival, reproduction,
1046 avoidance behavior and oxidative stress biomarkers in the earthworm *Octolasion cyaneum* exposed
1047 to glyphosate. *Bulletin of Environmental Contamination and Toxicology* 96(3), 314-319.
1048

- 1049 Samanta, P., Pal, S., Mukherjee, A.K., Ghosh, A.R., 2014. Biochemical effects of glyphosate based
1050 herbicide, Excel Mera 71 on enzyme activities of acetylcholinesterase (AChE), lipid peroxidation
1051 (LPO), catalase (CAT), glutathione-S-transferase (GST) and protein content on teleostean fishes.
1052 *Ecotoxicology and Environmental Safety* 107, 120-125.
- 1053
- 1054 Santos-Sánchez, N.F., Salas-Coronado, R., Hernández-Carlos, B., Villanueva-Cañongo, C., 2019.
1055 Shikimic acid pathway in biosynthesis of phenolic compounds. *InTechOpen*. Retrieved from
1056 [https://www.intechopen.com/books/plant-physiological-aspects-of-phenolic-compounds/shikimic-](https://www.intechopen.com/books/plant-physiological-aspects-of-phenolic-compounds/shikimic-acid-pathway-in-biosynthesis-of-phenolic-compounds)
1057 [acid-pathway-in-biosynthesis-of-phenolic-compounds](https://www.intechopen.com/books/plant-physiological-aspects-of-phenolic-compounds/shikimic-acid-pathway-in-biosynthesis-of-phenolic-compounds).
- 1058
- 1059 SAS, 2013. Base SAS 9.4 Procedures Guide: Statistical Procedures. SAS Institute Inc.
- 1060
- 1061 Saska, P., Skuhrovec, J., Lukas, J., Chi, H., Tuan, S.J., Honek, A., 2016. Treatment by glyphosate-
1062 based herbicide alters life history parameters of the rose-grain aphid *Metopolophium dirhodum*.
1063 *Scientific Reports* 6, 27801.
- 1064
- 1065 Schneider, M.I., Sanchez, N., Pineda, S., Chi, H., Ronco, A., 2009. Impact of glyphosate on the
1066 development, fertility and demography of *Chrysoperla externa* (Neuroptera: Chrysopidae):
1067 ecological approach. *Chemosphere* 76(10), 1451-1455.
- 1068
- 1069 Shilo, T., Zygier, L., Rubin, B., Wolf, S., Eizenberg, H., 2016. Mechanism of glyphosate control of
1070 *Phelipanche aegyptiaca*. *Planta* 244(5), 1095-1107.
- 1071
- 1072 Siehl, D.L., 1997. Inhibitors of EPSP synthase, glutamine synthetase and histidine synthesis. IOS
1073 Press, Amsterdam. Retrieved from
1074 [https://www.researchgate.net/publication/292667029_Inhibitors_of_EPSP_synthase_glutamine_syn-](https://www.researchgate.net/publication/292667029_Inhibitors_of_EPSP_synthase_glutamine_synthetase_and_histidine_synthesis)
1075 [thetase_and_histidine_synthesis](https://www.researchgate.net/publication/292667029_Inhibitors_of_EPSP_synthase_glutamine_synthetase_and_histidine_synthesis)
- 1076
- 1077 Sihtmäe, M., Blinova, I., Künnis-Beres, K., Kanarbik, L., Heinlaan, M., Kahru, A., 2013.
1078 Ecotoxicological effects of different glyphosate formulations. *Applied Soil Ecology* 72, 215-224.
- 1079
- 1080 Silva, V., Montanarella, L., Jones, A., Fernández-Ugalde, O., Mol, H.G.J., Ritsema, C.J., Geissen,
1081 V., 2018. Distribution of glyphosate and aminomethylphosphonic acid (AMPA) in agricultural
1082 topsoils of the European Union. *Science of the Total Environment* 621, 1352-1359.
- 1083
- 1084 Singh, R.J., 2002. Glutathione: A marker and antioxidant for aging. *Journal of Laboratory and*
1085 *Clinical Medicine* 140(6), 380-381.
- 1086
- 1087 Slaninová, A., Smutna, M., Modra, H., Svobodova, Z., 2009. A review: oxidative stress in fish
1088 induced by pesticides. *Neuroendocrinology Letters* 30, 2-12.
- 1089
- 1090 Smith, P.K., Krohn, R.I., Hermanson, G.T., Mallia, A.K., Gartner, F.H., Provenzano, M.D.,
1091 Fujimoto, E.K., Goeke, N.M., Olson, B.J., Klenk, D.C., 1985. Measurement of protein using
1092 bicinchoninic acid. *Analytical Biochemistry* 150(1), 76-85.
- 1093
- 1094 Steinrücken, H.C., Amrhein, N., 1980. The herbicide glyphosate is a potent inhibitor of 5-
1095 enolpyruvylshikimic acid-3-phosphate synthase. *Biochemical and Biophysical Research*
1096 *Communications* 94, 1207-1212.
- 1097

- 1098 Thompson, H.M., Levine, S.L., Doering, J., Norman, S., Manson, P., Sutton, P., von Merye, G.,
1099 2014. Evaluating Exposure and Potential Effects on Honeybee Brood (*Apis mellifera*) Development
1100 Using Glyphosate as an Example. *Integrated Environmental Assessment and Management* 10, 463-
1101 470.
- 1102
1103 Torretta, V., Katsoyiannis, A.I., Viotti, P., Rada, C.E., 2018. Critical Review of the Effects of
1104 Glyphosate Exposure to the Environment and Humans through the Food Supply Chain.
1105 *Sustainability* 10.
- 1106
1107 Tzin, V., Galili, G., 2010. New Insights into the Shikimate and Aromatic Amino Acids Biosynthesis
1108 Pathways in Plants. *Molecular Plant* 3, 956-972.
- 1109
1110 Uren Webster, T.M., Santos, E.M., 2015. Global transcriptomic profiling demonstrates induction of
1111 oxidative stress and of compensatory cellular stress responses in brown trout exposed to glyphosate
1112 and Roundup. *BMC Genomics* 16(32), 1254-1255.
- 1113
1114 Van Bruggen, A.H.C., He, M.M., Shin, K., Mai, V., Jeong, K.C., Finckh, M.R., Morris, J.G., 2018.
1115 Environmental and health effects of the herbicide glyphosate. *Science of the Total Environment*
1116 616-617, 255-268.
- 1117
1118 Vereecken, H., 2005. Mobility and leaching of glyphosate: a review. *Pest Management Science*
1119 61(12), 1139-1151.
- 1120
1121 Vuori, K.A., Lehtonen, K.K., Kanerva, M., Peltonen, H., Nikinmaa, M., Berezina, N.A., Boikova,
1122 E., 2015. Oxidative stress biomarkers in the copepod *Limnocalanus macrurus* from the northern
1123 Baltic Sea: effects of hydrographic factors and chemical contamination. *Marine Ecology Progress*
1124 *Series* 538, 131-144.
- 1125
1126 Vänninen, I, Worner, S., Huusela-Veistola, E., Tuovinen, T., Nissinen, A., Saikkonen, K., 2011.
1127 Recorded and potential alien invertebrate pests in Finnish agriculture and horticulture. *Agricultural*
1128 *and Food Science* 20(1), 96-114.
- 1129
1130 Walsh, B.D., 1865. The new potato bug and its natural history. *The Practical Entomology* 1, 1-4.
- 1131
1132 Ward, E., 1984. Suppression of metalaxyl activity by glyphosate: evidence that host defence
1133 mechanisms contribute to metalaxyl inhibition of *Phytophthora megasperma* f. sp. *glycinea* in
1134 soybeans. *Physiological Plant Pathology* 25(3), 381-386.
- 1135
1136 Woodburn, A.T., 2000. Glyphosate: production, pricing and use worldwide. *Pest Management*
1137 *Science* 56(4), 309-312.
- 1138
1139 Yang, D.-B., Xu, Y.-C., Wang, D.-H., Speakman, J.R., 2013. Effects of reproduction on immuno-
1140 suppression and oxidative damage, and hence support or otherwise for their roles as mechanisms
1141 underpinning life history trade-offs, are tissue and assay dependent. *Journal of Experimental*
1142 *Biology* 216, 4242-4250.
- 1143
1144 Zobiolo, L.H.S., Kremer, R.J., de Oliveira Jr., R.S., Constantin, J., 2012. Glyphosate effects on
1145 photosynthesis, nutrient accumulation, and nodulation in glyphosate-resistant soybean. *Journal of*
1146 *Plant Nutrition and Soil Science* 175(2), 319-330.
- 1147

1148 **Appendices:**

Table A1. Mean (\pm 95% CI) activities of oxidative status parameters: glutathione-S-transferase (GST), glutathione oxidase (GPx), glutathione reductase (GR), catalase (CAT), superoxide dismutase (SOD), total glutathione (tGSH), ratio of reduced and oxidized glutathione (GSH:GSSG), and lipid hydroperoxides (LHP) in control and GBH treatment groups of larval and adult Colorado potato beetles (*L. decemlineata*).

Biomarker	Larvae				Adults			
	Control		GBH		Control		GBH	
	Mean (\pm 95% CI)	n	Mean (\pm 95% CI)	n	Mean (\pm 95% CI)	n	Mean (\pm 95% CI)	n
GST (nmol/min/mg)	14.62 (13.26-15.97)	34	17.28 (14.65-19.91)	34	35.76 (32.12-39.41)	32	35.15 (31.60-38.71)	32
GPx (nmol/min/mg)	5.31 (4.90-5.72)	34	5.59 (5.06-6.13)	34	2.71 (1.67-3.75)	30	3.17 (1.75-4.58)	31
GR (nmol/min/mg)	4.93 (3.94-5.92)	33	5.37 (3.86-6.88)	33	4.03 (3.30-4.77)	32	3.78 (3.09-4.46)	32
CAT (μ mol/min/mg)	143.04 (134.46-151.63)	34	165.80 (154.98-176.63)	31	116.90 (107.74-126.06)	32	103.14 (90.67-115.61)	32
SOD (inhibition %)	73.70 (71.65-75.74)	34	77.42 (75.31-79.54)	34	80.94 (78.84-83.04)	32	78.12 (75.71-80.53)	32
tGSH (μ mol/mg)	11.23 (8.78-13.68)	20	20.77 (17.55-23.99)	23	41.88 (35.88-47.89)	31	40.35 (33.98-46.72)	25
GSH:GSSG (ratio)	0.45 (0.082-0.83)	20	0.62 (0.23-1.01)	23	3.51 (2.39-4.62)	29	4.37 (1.84-6.89)	25
LHP (nmol/mg bm)	0.57 (0.11-1.04)	16	0.40 (-0.01-0.82)	17	0.018 (0.014-0.023)	27	0.017 (0.014-0.020)	30

1149

1150

Table A2 A. Spearman correlation coefficients (r^2 , p-value, n) between the potato glycoalkaloids (α -solanine and α -chaconine) and oxidative status biomarkers glutathione-S-transferase (GST), glutathione oxidase (GPx), glutathione reductase (GR), catalase (CAT), superoxide dismutase (SOD), total glutathione (tGSH), ratio of reduced and oxidized glutathione (GSH:GSSG), lipid hydroperoxides (LHP) and body mass (g) in the Colorado potato beetle larvae (*L. decemlineata*) in the GBH treatment.

		GST	GPx	GR	CAT	SOD	tGSH	GSH:GSSG	LHP	Body mass
α-solanine (1)	r^2	0.111	-0.243	-0.163	-0.517	-0.126	-0.326	-0.041	-0.476	-0.387
	p	0.707	0.402	0.594	0.070	0.668	0.328	0.904	0.233	0.171
	n	14	14	13	13	14	11	11	8	14
α-solanine (2)	r^2	-0.558	0.053	-0.202	-0.694	-0.268	-0.436	-0.592	-0.167	0.144
	p	0.038	0.857	0.508	0.009	0.355	0.180	0.055	0.693	0.624
	n	14	14	13	13	14	11	11	8	14
α-chaconine (1)	r^2	-0.513	-0.226	-0.147	-0.606	-0.285	-0.454	-0.537	-0.286	0.002
	p	0.06	0.438	0.632	0.028	0.323	0.161	0.089	0.493	0.994
	n	14	14	13	13	14	11	11	8	14
α-chaconine (2)	r^2	-0.593	0.199	-0.091	-0.628	-0.215	-0.087	-0.500	-0.048	0.400
	p	0.025	0.495	0.767	0.022	0.461	0.799	0.117	0.911	0.156
	n	14	14	13	13	14	11	11	8	14

1151

Table A2 B. Spearman correlation coefficients (r^2 , p-value, n) between the potato glycoalkaloids (α -solanine and α -chaconine) and oxidative status biomarkers glutathione-S-transferase (GST), glutathione oxidase (GPx), glutathione reductase (GR), catalase (CAT), superoxide dismutase (SOD), total glutathione (tGSH), ratio of reduced and oxidized glutathione (GSH:GSSG), lipid hydroperoxides (LHP) and body mass (g) in the Colorado potato beetle larvae (*L. decemlineata*) in the control treatment.

		GST	GP	GR	CAT	SOD	tGSH	GSH: GSSG	LHP	Body mass
α-solanine (1)	r^2	-0.193	0.067	-0.034	0.269	0.168	0.154	0.410	-0.257	0.269
	p	0.618	0.864	0.932	0.484	0.666	0.805	0.493	0.623	0.484
	n	9	9	9	9	9	5	5	6	9
α-solanine (2)	r^2	0.193	0.185	-0.135	-0.824	-0.572	0.667	-0.205	-0.371	-0.303
	p	0.618	0.634	0.730	0.006	0.108	0.219	0.741	0.469	0.429
	n	9	9	9	9	9	5	5	6	9
α-chaconine (1)	r^2	-0.126	0.252	-0.118	0.017	0.168	0.154	0.410	-0.257	0.168
	p	0.747	0.513	0.763	0.966	0.666	0.805	0.493	0.623	0.666
	n	9	9	9	9	9	5	5	6	9
α-chaconine (2)	r^2	0.261	0.387	0.151	-0.656	-0.454	0.667	-0.205	0.029	-0.437
	p	0.498	0.304	0.698	0.055	0.220	0.219	0.741	0.957	0.240
	n	9	9	9	9	9	5	5	6	9

1152

1153

Table A2 C. Spearman correlation coefficients (r^2 , p-value, n) between the potato glycoalkaloids (α -solanine and α -chaconine) and oxidative status biomarkers glutathione-S-transferase (GST), glutathione oxidase (GPx), glutathione reductase (GR), catalase (CAT), superoxide dismutase (SOD), total glutathione (tGSH), ratio of reduced and oxidized glutathione (GSH:GSSG), lipid hydroperoxides (LHP) and body mass (g) in the Colorado potato beetle adults (*L. decemlineata*) in the GBH treatment.

		GST	GPx	GR	CAT	SOD	tGSH	GSH: GSSG	LHP	Body mass
α-solanine (1)	r^2	-0.062	-0.061	0.064	-0.021	0.054	-0.050	-0.177	-0.102	0.341
	p	0.807	0.810	0.801	0.932	0.832	0.859	0.528	0.687	0.167
	n	18	18	18	18	18	15	15	18	18
α-solanine (2)	r^2	0.068	0.131	0.019	0.199	0.180	0.032	-0.134	0.331	0.250
	p	0.788	0.604	0.942	0.428	0.476	0.909	0.634	0.179	0.317
	n	18	18	18	18	18	15	15	18	18
α-chaconine (1)	r^2	0.165	0.049	0.015	-0.018	-0.025	-0.093	-0.120	0.084	-0.066
	p	0.512	0.848	0.955	0.945	0.922	0.742	0.671	0.741	0.795
	n	18	18	18	18	18	15	15	18	18
α-chaconine (2)	r^2	0.235	0.179	-0.079	0.129	-0.006	0.004	-0.216	0.206	0.145
	p	0.347	0.478	0.757	0.610	0.981	0.990	0.439	0.413	0.567
	n	18	18	18	18	18	15	15	18	18

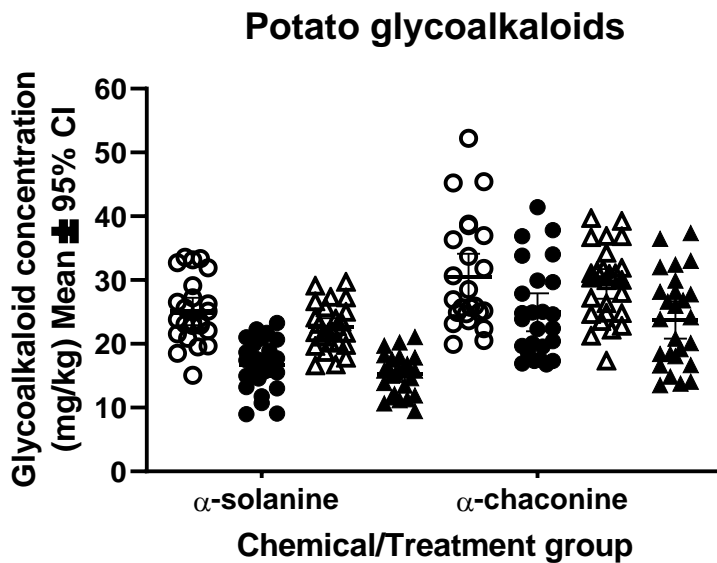
1154

1155

Table A2 D. Spearman correlation coefficients (r^2 , p-value, n) between the potato glycoalkaloids (α -solanine and α -chaconine) and oxidative status biomarkers glutathione-S-transferase (GST), glutathione oxidase (GPx), glutathione reductase (GR), catalase (CAT), superoxide dismutase (SOD), total glutathione (tGSH), ratio of reduced and oxidized glutathione (GSH:GSSG), lipid hydroperoxides (LHP) and body mass (g) in the Colorado potato beetle adults (*L. decemlineata*) in the control treatment.

		GST	GPx	GR	CAT	SOD	tGSH	GSH:GSSG	LHP	Body mass
α -solanine (1)	r^2	-0.385	-0.005	-0.218	-0.096	0.039	0.010	-0.282	-0.558	0.437
	p	0.127	0.985	0.400	0.715	0.881	0.970	0.273	0.031	0.070
	n	17	17	17	17	17	17	17	15	18
α -solanine (2)	r^2	-0.128	-0.135	-0.306	-0.230	-0.326	-0.289	-0.24	0.075	0.385
	p	0.626	0.606	0.232	0.374	0.202	0.260	0.353	0.790	0.115
	n	17	17	17	17	17	17	17	15	18
α -chaconine (1)	r^2	-0.299	-0.164	-0.015	-0.341	0.005	0.159	-0.326	-0.329	0.270
	p	0.244	0.529	0.955	0.181	0.985	0.541	0.202	0.231	0.280
	n	17	17	17	17	17	17	17	15	18
α -chaconine (2)	r^2	-0.103	-0.174	0.034	-0.279	-0.081	-0.015	-0.123	-0.021	-0.038
	p	0.694	0.504	0.896	0.277	0.758	0.955	0.639	0.940	0.880
	n	17	17	17	17	17	17	17	15	18

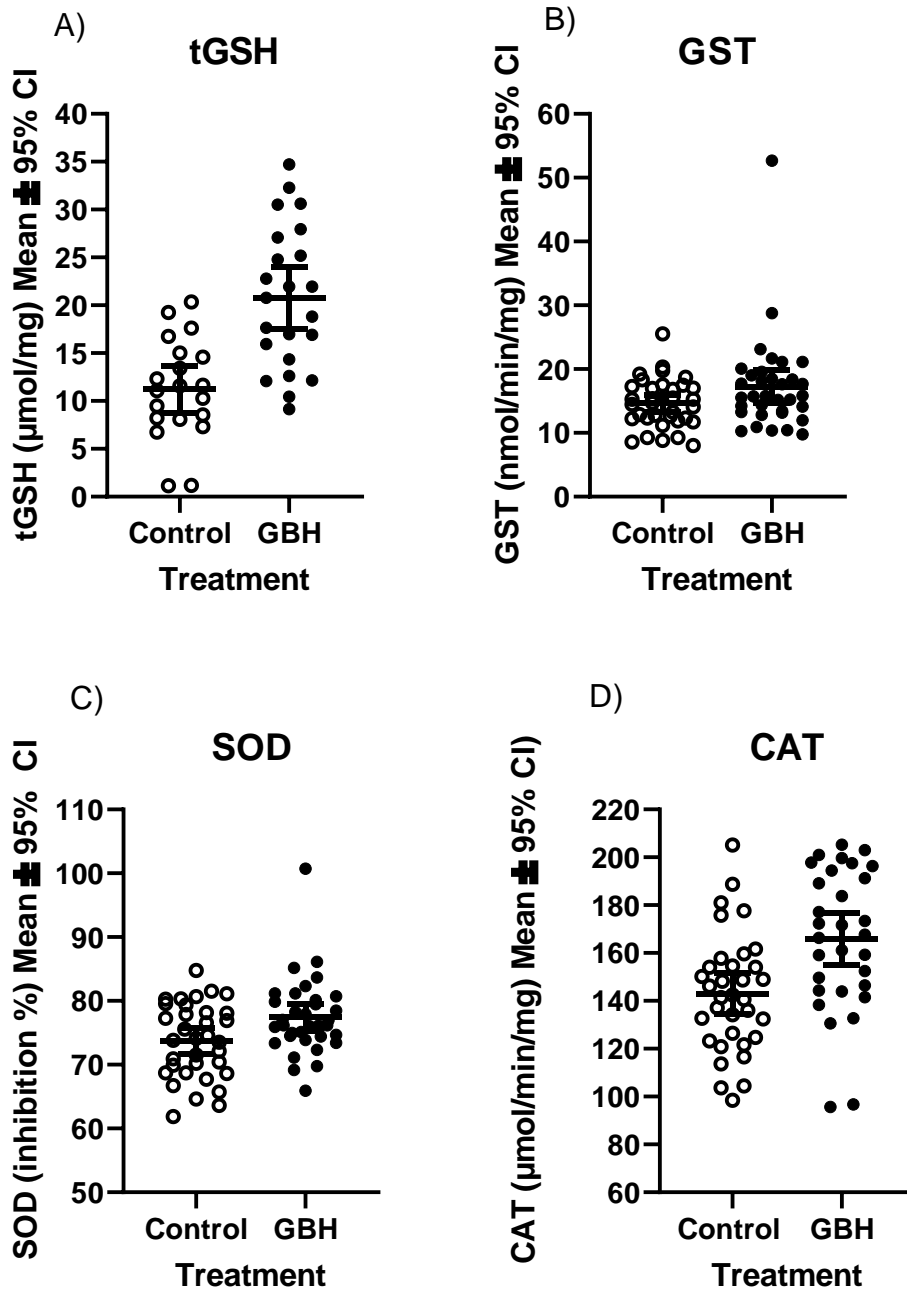
1156



1157

1158 **Figure A1.** Potato glycoalkaloid (α -solanine and α -chaconine) concentrations (raw data mean \pm 95%
 1159 CI) between the treatment groups (control=circle, GBH treatment=triangle) at two measurement
 1160 points (measurement 1=white, measurement 2=black).

1161



1162

1163

1164 **Figure A2.** Variation in A) total glutathione (tGSH) concentration, B) glutathione-S-transferase
 1165 (GST), C) superoxide dismutase (SOD), and D) catalase (CAT) activity in larvae of the Colorado
 1166 potato beetle (*L. decemlineata*) between treatment groups (control=white circle, GBH
 1167 treatment=black circle). The dots represent the raw data (mean \pm 95% CI).

1168